

Article

Foliage biomass of the genera *Picea* spp. and *Quercus* spp. in winter temperature and annual precipitation gradients: Inter-genera paradox in the forests of Eurasia

V. A. Usoltsev^{2,3}, H. Lin^{1,4,5,6}, S.O. R. Shobairi¹, I. S. Tsepordey³, Z. Ye^{1,4,5,6}

¹Research Center of Forestry Remote Sensing & Information Engineering, Central South University of Forestry and Technology, Changsha 410004, China

²Ural State Forest Engineering University, Faculty of Forestry, Sibirskiy Trakt, 37, 620100 Yekaterinburg, Russia

³Botanical Garden of Ural Branch of RAS, Department of Forest Productivity, ul. 8 Marta, 202a, 620144 Yekaterinburg, Russia

⁴Key Laboratory of Forestry Remote Sensing Based Big Data & Ecological Security for Hunan Province, Changsha 410004, China

⁵Key Laboratory of State Forestry Administration on Forest Resources Management and Monitoring in Southern Area, Changsha 410004, China

⁶Changsha Changchang Forestry Technology Consulting Co., Ltd., Changsha 410004, China

E-mails: Usoltsev50@mail.ru, linhui@csuft.edu.cn, Omidshobeyri214@gmail.com, common@botgard.uran.ru, 1053460198@qq.com

Received 19 April 2022; Accepted 25 May 2022; Published online 30 May 2022; Published 1 September 2022



Abstract

Current climate change make it increasingly important to assess the response of forest cover biomass to this change, which in turn determines the possibility of climate stabilization by reducing atmospheric CO₂. Climate change projections can be linked to significant changes in water deficits and natural disturbance regimes in forest ecosystems across many parts of the world. Since the carbon-depositing potential of a plant community is determined by the biomass of assimilating organs, we focused in this article on comparative analysis of foliage biomass in the dark-coniferous genus *Picea* spp. and of leaved genus *Quercus* spp. in gradients of annual precipitation and winter temperature on the territory of Eurasia. The database for modelling involves 870 and 570 sample plots for spruce and oak correspondingly. When studying changes in the biomass of spruce and oak foliage in the hydrothermal gradients of Eurasia, we obtained statistically reliable, but surprising and paradoxical results. When the temperature increases by 1°C, the most increase in spruce foliage biomass occurs in cold regions with insufficient moisture supply, while oak leaf biomass, on the contrary, has the largest percentage of its decrease. In cold regions, as the transition from dry conditions to wet ones, the percentage of foliage mass in spruce decreases, and in oak increases. When the precipitation increases by 100 mm, there is the opposite patterns for spruce and oak: in warm regions, spruce has a decrease in the percentage of foliage mass, and oak, on the contrary, its increase. In the cold regions, these patterns change to the opposite: the spruce negative percentage changes to positive, and the oak positive percentage changes to negative. This phenomenon seems to be related to the positive ratio of photosynthesis and respiration in the spruce during the winter period when the oak does not have photosynthesis, but only respiration. It seems to be related to the fundamental difference between the winter physiology of evergreen and deciduous species, namely the ability of the former to assimilate atmospheric carbon dioxide and prolong the assimilation process beyond the vegetation period, which is usual for deciduous species. The development of such models for the main forest-forming species of Eurasia allow us to predict changes in the foliage productivity of the forest cover of Eurasia in relation to climate change.

Keywords hydrothermal gradients; foliage biomass; foliage efficiency; regression models; biomass equations; mean January temperature; annual precipitation.

Computational Ecology and Software
ISSN 2220-721X
URL: <http://www.iaees.org/publications/journals/ces/online-version.asp>
RSS: <http://www.iaees.org/publications/journals/ces/rss.xml>
E-mail: ces@iaees.org
Editor-in-Chief: WenJun Zhang
Publisher: International Academy of Ecology and Environmental Sciences

1 Introduction

The basis for the functioning of the biosphere is its biological productivity, and the life of humans and all heterotrophic organisms depends on it, but only in the last 50-60 years the idea of the biosphere's productive capacity has gone beyond "enlightened guesses" and received some quantitative confirmation (Whittaker et al., 1975). The main pool of recent publications on forest biomass is related to the assessment of possible changes in vegetation cover under the influence of climate, but most of them consider global estimates without taking into account the species composition, age and morphological structure, which do not answer the question of how the biological productivity of the planet's forest cover will change if the temperature changes by a certain number of degrees or precipitation changes by a certain number of mm (Anderson et al., 2006; Keeling and Phillips, 2007; Eggers et al., 2008; Huston and Wolverton, 2009; Shuman and Shugart, 2009; Poudel et al., 2011; Han et al., 2018; Zeller et al. 2018).

At the level of forest communities, numerous studies of biomass structure are related to climate variables (Lieth, 1974; Monserud et al., 1996; Ni et al., 2001; Lapenis et al., 2005; Shuman et al., 2011; Stegen et al., 2011; Poudel et al., 2011; Pretzsch et al., 2012; Lewis et al., 2013; Berner et al., 2013; Frank et al., 2015; Schaphoff et al., 2016; Fang et al., 2016; Dong and Liu, 2017; Wang et al., 2017; Gu et al., 2017; Gustafson et al., 2017; Duan et al., 2018; Paquette et al., 2018; Sippel et al., 2018; Teets et al., 2018; Wang et al., 2018; Sperry et al., 2019; Khan et al., 2019a, b; Dolezal et al., 2021). Contradictory results were obtained even within a single region, while the influence of the same climatic factors on the biological production of stands of certain tree species (genera) in the Trans-Eurasian climatic gradients of temperature and precipitation is still unknown, since the available regional information is fragmentary and contradictory (Strömngren and Linder, 2002; Wilmking et al., 2004; Huang et al. 2010; Stegen et al., 2011; Fu et al., 2017; Forrester et al. 2017; Reichstein and Carvalhais, 2019).

However, there is the common agreement that vegetation responses to climate change are species-specific (Schulze and Mooney, 1994; Spathelf et al., 2018). It is well known that tree species (genera) occupy specific ecological niches and react differently to short-term stressful situations (Kusmin, 1930; Vassiljev, 1931; Gower and Richards, 1990; Suvorova et al., 1999; Suvorova, 2009; Sudachkova et al., 2012), and on long-term climate change too, but in the latter cases the plant trait responses are not only species (Pucko et al., 2011; Buckley and Kingsolver, 2012; Parmesan and Hanley, 2015; Gray and Brady, 2016; Baruah et al., 2017) but also community-specific ones (Dymond et al., 2016; Baruah et al., 2017; Noce et al., 2019; Ye et al., 2022).

It is believed that coniferous and deciduous species of the temperate zone do not differ in net primary production (Schulze, 1982; Reich and Bolstad, 2001). Since these two different functional types of forests are adapted to specific forest conditions and may occupy different ecological niches, the same type of forest conditions may, for example, be more suitable for coniferous than for deciduous, and vice versa (Givnish, 2002). Coniferous species are more adapted to harsh habitats that have a slow cycle of substances, which

allows them to limit the amount of assimilates entering the needles. Deciduous species are often more adapted to a mild climate that provides an accelerated cycle of nutrients, which is necessary for deciduous species to meet the increased cost of annual leaf replacement (Schulze, 1982). Nevertheless, net primary production was consistently larger for deciduous than for evergreen boreal forests in each of the major boreal regions, especially for boreal forests in Alaska (Gower et al., 2001).

In comparison with these uncertainties and contradictions, we obtained a much more surprising and paradoxical result when studying changes in the foliage biomass of larch and pine in the hydrothermal gradients of Eurasia, which is described in this article.

The genus *Picea* spp. includes about 45 species, but a relatively intensive hybridisation complicates even their simple description. Usually Mayr's classification scheme was applied, which splits the genus into three sections – *Morinda*, *Casieta* and *Omorica* (Bobrov, 1978). Common spruce (*P. abies* [L.]Karst.), Siberian spruce (*P. obovata* Ldb.), Caucasian spruce (*P. orientalis* [L.] Link), Schrenk's spruce (*P. schrenkiana* F. et M.) and Jezo spruce (*P. jezoensis* [Siebold and Zucc.] Carrière) are the most widely distributed in Eurasia. The common spruce area is fragmented and set before with three local areas – Alpine, Carpathian and Baltic – that independently developed in the postglacial time. The distribution of common spruce to the south of Western Europe and almost to the Arctic Ocean shows that its climate adaptation has a wide range. It is believed that *Picea* spp. has a mountain origin, which is confirmed by the surface type of root system and the unique feature of natural renewal, most successful on detritus - rotting remnants of fallen trees, and in different types of forests and in different climatic conditions within its natural range (Usoltsev, 2019).

The genus *Quercus* spp. includes about 600 species in the temperate and tropical zone of the Northern Hemisphere, a distinguishing dependence of oak forests on climate is their affiliation with the humid regions with reduced continentality. In the temperate climate of Western Europe, they range from the Mediterranean Sea and deep into the north whereas in more continental conditions in the East European Plain geographically they are localized in a relatively narrow band. In Russia the oak forests are less than 1% of the forested area. So, the genus *Quercus* is divided into the western and eastern parts of the Eurasian continent, developed in the Tertiary period and varied by biological and especially ecological characteristics. There deciduous and ever-green trees, slow and fast growing at the young age, large trees and sub-shrubs, drought resistant and long-term water logging tolerated species. However, there is one common feature for all these species, which is the affiliation with the humid region with the reduced continentality (Usoltsev, 2019).

The specificity of the response of tree species to climate change may be related to their interaction in different forest growing conditions (Lavrinenko, 1965). The question of relationship between spruce and oak in the forest-steppe of Russia has its own history. It is known the theoretical dispute between botanists (Korzhinsky, 1888) and foresters (Morozov, 1949) broking out on the relationship between oak and spruce in this zone. Geobotanists argued that the spruce that settled in the oak forests will inevitably displace the oak. Foresters were of the opinion that after the introduction of spruce under the canopy of oak forests, spruce is doomed to extinction (Morozov, 1949), which was later confirmed by Yartukin (1968). Khitrovo (1907) was the most categorical, and gave an example when the spruce and oak ecosystems have been facing each other for thousands of years, separated by the Volga only, without any noticeable tendency to displace the oak from the spruce and vice versa.

To avoid uncertain results mentioned above, a transcontinental level of analysis was chosen in our study, using the opportunities and prospects that are opening up today through "big data" technologies (Niu et al., 2020). The purpose of this study was to show how much the foliage biomass of the genus *Picea* spp. can change with a possible increase in temperature by 1°C at constant precipitation and with a possible increase in

precipitation by 100 mm per year at constant temperature, and to compare the results with similar data for the genus *Quercus* spp. in Eurasia. Leaf biomass can provide information about plant investments in leaf tissue and correlates with functional types of woody plants, such as the specific leaf area that can be used in modeling plant water consumption (Mencuccini et al., 2019). Since the productive and carbon-depositing potential of a plant community is determined by the biomass of assimilating organs, we have focused in this article not on the total biomass of forest stands, but only on the mass of foliage.

2 Material and Methods

Since the response to climate change differs between pure and mixed stands (Paquette et al., 2018), in our comparative study we used harvest data only from pure stands that were selected from our database (Usoltsev, 2020) in a number of 870 and 570 sample plots with estimations of spruce and oak forest stand biomass (t/ha) is extracted. Their fragment one can see in Table 1.

Table 1 A fragment of the original matrix of experimental data*.

<i>Picea</i> forests						<i>Quercus</i> forests					
<i>A</i>	<i>N</i>	<i>V</i>	<i>P_f</i>	<i>T_m</i>	<i>PR_m</i>	<i>A</i>	<i>N</i>	<i>V</i>	<i>P_f</i>	<i>T_m</i>	<i>PR_m</i>
17	5.000	76	24.0	-3	826	70	0.880	277	5.70	-2	826
10	5.500	10	8.00	-3	826	70	0.572	180	4.10	-2	826
20	4.128	58	19.7	-3	826	80	0.490	450	7.23	-2	826
115	0.300	523	12.66	-2	600	60	0.820	328	5.37	-7	570
116	0.300	561	12.83	-2	600	70	0.628	376	5.71	-7	570
117	0.284	541	12.32	-2	600	80	0.498	417	6.04	-7	570
30	17.72	52	5.91	-14	570	55	1.486	129	3.46	3	635
40	7.267	82	8.38	-14	570	20	7.000	73	3.40	3	635
50	3.775	111	10.8	-14	570	28	7.000	74	2.50	3	635
46	0.756	252	14.7	-5	826	47	0.618	164	1.64	5	826
45	0.914	193	16.9	-5	826	44	0.890	117	2.17	5	826
39	2.240	368	23.9	0	890	21	10.102	44	3.38	5	826
30	1.072	269	24.6	0	890	100	0.500	207	16.1	10	1140
92	0.568	614	9.20	0	635	100	0.570	290	21.6	10	1140
85	0.475	900	20.8	0	635	100	0.725	325	23.5	10	1140
50	1.931	144	1.21	-20	317	20	15.47	59	4.74	-15	570
70	1.142	257	1.80	-20	317	40	1.240	133	3.19	-15	570
90	0.806	344	1.97	-20	317	60	0.634	202	3.57	-15	570
100	1.670	276	9.10	-30	570	80	0.615	165	1.00	-6	570
210	0.247	91	2.70	-30	570	55	1.648	346	6.01	-6	570
220	1.000	221	7.30	-30	570	59	1.723	378	6.84	-6	570
48	1.082	113	11.51	5	1140	80	4.000	150	6.00	3	890
214	0.401	637	7.69	5	1140	80	3.700	170	7.60	3	890
317	0.405	1143	10.0	5	1140	80	5.200	195	7.10	3	890
164	0.501	472	8.81	5	1140	13	10.30	124	5.30	0	826

*Designations here and further: *A* = stand age, yrs; *V* = stem volume, m³/ha; *N* = tree density, 1000/ha; *P_f* = foliage biomass, t per ha; *T_m* = mean January temperature, °C; *PR_m* = mean annual precipitation, mm.

Distribution of data collected from different sites is shown in Fig. 1. We can see that the coverage of the territory of Eurasia by *Picea* and *Quercus* is quite different, apparently, due to the difference in their ecological niches. Each sample plot, in which the biomass of the forest stands was estimated, is positioned in accordance to January mean annual temperature isolines and to mean annual precipitation isolines, and the initial data matrix is compiled in which the values of biomass components and of stand taxation characteristics are mated with corresponding values of mean January temperature and precipitation taken from World Weather Maps (2007). The matrix is then subjected to regression analysis.

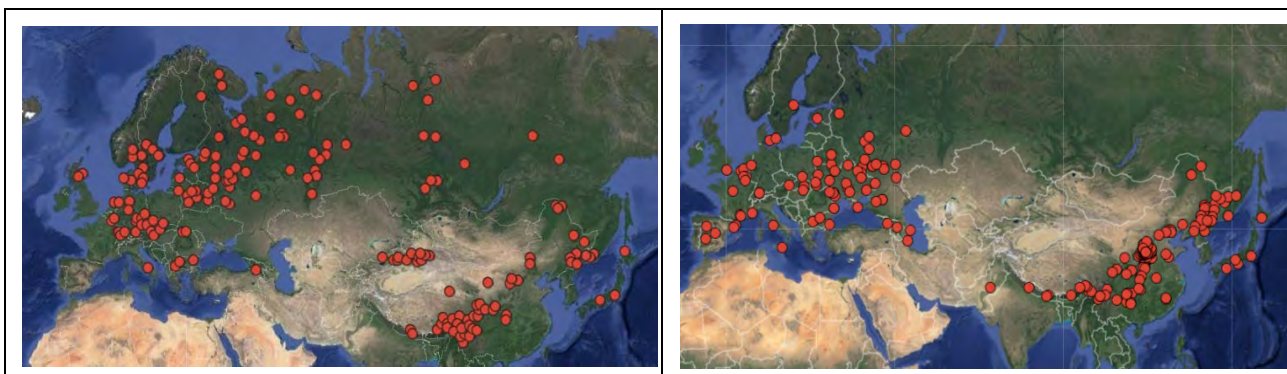


Fig. 1 The distribution of 870 sampling sites of *Picea* (on the left) and 570 sampling sites of *Quercus* (on the right) on the territory of Eurasia.

The question may arise why modelling was performed at a level of genera, and not for individual spruce and oak species. If we adhere to the concept of species-specific responses of forest biomass to changes in the main climatic characteristics, then when we reach the transcontinental level, we are faced with the obvious fact that no species grows throughout the continent, precisely because of regional climate differences. Moving from refuges under the influence of geological processes and climate changes, the particular species adapted to changing environmental conditions, forming a series of vicariate species within a genus (Hultén, 1937; Tolmachev, 1962; Chernyshev, 1974). This gives grounds for analyzing the response of tree species to changes in climate characteristics, to combine them into one climate-dependent set within the entire genus, since differences in ecological and physiological properties of different species of the genus, for example, *P. abies* (L.) H. Karst. vs. *P. obovata* L. vs. *P. koraiensis* Nakai are derived from regional climatic features.

It is well known that when estimating stem biomass growth by using the annual ring width, the greatest contribution to explaining its variability being made by summer temperature. Moreover, it was established that this relationship is positive with the maximum intra-annual temperature and negative with the minimum and average annual temperature (Khan et al., 2019). With an inter-annual time step, the predominant influence of summer temperature is quite normal (Levanič et al., 2015; Zubairov et al., 2018; Bocharnikov, 2019). But against the background of long-term climatic shifts for decades, the prevailing influence is acquired by winter temperatures (Toromani and Bojaxhi, 2010; Bijak, 2010; Morley et al., 2017), having in mind that winter temperatures in the Northern Hemisphere are increased faster than summer ones during the 20th century (Emanuel et al., 1985; Folland et al., 2001; Laing and Binyamin, 2013; Felton et al., 2016). This phenomenon may be associated with the shift of the earth's magnetic pole towards Siberia (Olsen and Manda, 2007).

In terms of regression analysis, a weak temporal trend of summer temperatures compared to a steep trend of winter ones means a smaller regression slope and a worse ratio of residual variance to the total

variance explained by this regression. Obviously, taking the mean winter temperature as one of the independent variables, we get a more reliable dependence having the higher predictive ability. To ensure the maximum stability of the model, each of the selected factors (independent variables) should be presented in the maximum range of its variation (Usoltsev, 2003). In our example, mean January temperatures ranged from -40°C in the forest-tundra of the North–Eastern Siberia to $+10^{\circ}\text{C}$ in the subtropics of China. We compiled precipitation data ranging from 190 mm in permafrost regions of North-Eastern Siberia to 1,140 mm in South of China.

The matrix was used as a source of data in the subsequent regression analysis. It is well known, the biomass of a stand represented by a particular tree species is primarily determined by its age and morphological (taxation) structure, i.e., a set of characteristics such as age, mean height, mean diameter at breast height, the basal area, and the volume stock, which are interrelated. The problem of multicollinearity arises in empirical modeling of biomass. One of the solutions to the problem is to harmonize the system by constructing recursive (recurrent, related) equations for the different mensuration indices and foliage biomass, in which the dependent variable of the previous equation is included as one of the independent variables of the subsequent one (Draper and Smith, 1966; Usoltsev, 1990; Zhao et al., 2022). This approach has one of the methods for model harmonization (Jacobs and Cunia, 1980), provides a multivariate conditionality of factors that provide flexibility and universality of the regression system describing the dynamics of biomass of stands.

When choosing the structure of the regression model, we adhered to the concept that there is only one definite variant of stand biomass structure corresponding to a given structure of taxonomic parameters (morphological structure) of a tree stand (Usoltsev, 2007). The initial structure of the model included the main mass-forming indices of stands – age, stem volume, tree density, mean diameter and mean height. Mean diameter, and mean height were excluded in the process of the regression analysis as these factors were not statistically significant. The final structure of the recursive system of equations included only those mass-forming indices that were statistically significant for foliage biomass component:

$$\begin{aligned} \ln N &= f \{ \ln A, \ln(Tm+M), \ln PRm, [\ln(Tm+M)] \cdot (\ln PRm) \} \rightarrow \\ \rightarrow \ln V &= f \{ \ln A, \ln N, (\ln A)(\ln N), \ln(Tm+M), \ln PRm, [\ln(Tm+M)] \cdot (\ln PRm) \} \rightarrow \\ \rightarrow \ln Pi &= f \{ \ln A, \ln V, \ln N, (\ln A)(\ln N), \ln(Tm+M), \ln PRm, [\ln(Tm+M)] \cdot (\ln PRm) \}. \end{aligned} \quad (1)$$

3 Results and Discussion

The calculated equations (1) for foliage biomass are presented in Table 2. Because mean January temperature in northern part of Eurasia has negative values, the corresponding independent variable was increased by M ($Tm+M$) to enable a logarithmic transformation according to (Baskerville, 1972), where $M = 40$ for *Picea* and $M = 30$ for *Quercus*. The synergism $(\ln A) \cdot (\ln N)$ was introduced in the model to account for the decrease in the tree density with age and its effect on the foliage biomass. All the necessary calculations were carried out in Statgraphics software (<http://www.statgraphics.com/>).

First of all, we should pay attention to the signs for climate variables in the equations for spruce and oak (Table 2). In the equations for the foliage biomass, the regression coefficients for the variables: temperature, precipitation and their combined effect (synergy) are equal to $+3.9064$, $+1.8125$ and -0.6126 for spruce, and to -3.0578 , -0.8465 and $+0.4607$ for oak correspondingly, i.e. we have mirror patterns for the foliage biomass of spruce and oak. These regression coefficients are significant at the probability level of 0.99 for spruce, and of 0.95 in two climate variables out of the available three ones for oak, and the corresponding values of t-Student are $+3.45$, $+2.95$, and -3.37 for spruce foliage, and -2.05 , -1.14 , and $+2.00$ for oak foliage.

The proportions of the contribution of independent variables to the explanation of the variability of the forest biomass in equations (1) are shown in Table 3. We can see that mass-forming variables explain in averaged about 84, and 70% of the variability of foliage biomass in *Picea*, and *Quercus* correspondingly, including 54, and 29% of the contribution from the stem volume. Climate variables explain about 16, and 30% of the foliage biomass variability in *Picea*, and *Quercus* correspondingly, i.e. less than about 5.1, and 2.2 times as much as mass-forming variables in *Picea*, and *Quercus* correspondingly.

When tabulating a multifactorial biomass model that includes age, number of trees and volume stock as independent variables, a problem arises. This problem consists in the fact that we can insert only the desired age values into the model, and the values of the other two independent variables are unknown to us. The recursive principle of model construction gives us a solution to the problem by sequentially tabulating the equations. The recursive system of equations (1) was tabulated in the sequence indicated by the arrows: first, the values of tree density according to the given values of age, temperature and precipitation were obtained, then the volume stock according to the given values of age, temperature and rainfall and the calculated values of the tree density were obtained, and finally, the values of the foliage biomass according to the given values of age, temperature and precipitation, the calculated values of tree density and the volume stock were obtained. From the resulting tables, the values of the foliage biomass for the age of maturity (100 years for spruce and for oak) are taken, and 3-D graphs showing statistically significant transcontinental trends of biomass in temperature and precipitation gradients are designed (Fig. 2).

Table 2 Characteristic of equations (1) for forest stands of the *Picea*, and *Quercus* in Eurasia.

$\ln(Y)^{(1)}$	$a_0^{(2)}$	$\ln A$	$\ln V$	$\ln N$	$(\ln A) \cdot (\ln N)$	$\ln(Tm+M)$	$\ln PRm$	$[\ln(Tm+M)] \cdot (\ln PRm)$	$adjR^{2(3)}$	$SE^{(4)}$
<i>Picea</i>										
$\ln(N)$	-11.5190	-1.0091	-	-	-	5.2191	3.0637	-0.9733	0.569	0.61
$\ln(V)$	-11.2252	0.8455	-	-1.1457	0.2699	-2.9798	-2.4759	0.7765	0.642	0.59
$\ln(Pf)$	-10.8564	-0.3423	0.5893	-0.2302	0.1047	3.9064	1.8125	-0.6126	0.669	0.42
<i>Quercus</i>										
$\ln(N)$	8.7044	-1.2007	-	-	-	-1.2007	-0.9147	0.2921	0.618	0.59
$\ln(V)$	-5.2849	0.9559	-	-0.2394	0.0600	3.2725	0.8354	-0.4472	0.579	0.52
$\ln(Pf)$	5.3121	0.2723	0.1575	-0.0081	0.0428	-3.0578	-0.8465	0.4607	0.403	0.40

⁽¹⁾Dependent variables;

⁽²⁾The constant corrected for logarithmic retransformation by Baskerville (1972);

⁽³⁾ $adjR^2$ – determination coefficient adjusted for the number of variables;

⁽⁴⁾ SE – standard error of the equations.

Table 3 Contribution of independent variables of equations (1) to the explanation of variability of dependent variables (%).

$\ln(Y)$	Independent variables								
	$\ln A(I)$	$\ln V(II)$	$\ln N(III)$	$(\ln A) \cdot (\ln N)(IV)$	$I+II+III+IV$	$\ln(Tm+M)(V)$	$\ln PRm(VI)$	$[\ln(Tm+M)] \cdot (\ln PRm)(VII)$	$V+VI+VII$
<i>Picea</i>									
$\ln(Pf)$	15.1	53.9	5.2	9.5	83.7	5.8	4.9	5.6	16.3
<i>Quercus</i>									
$\ln(Pf)$	29.7	28.7	0.5	10.7	69.6	12.0	6.7	11.7	30.4

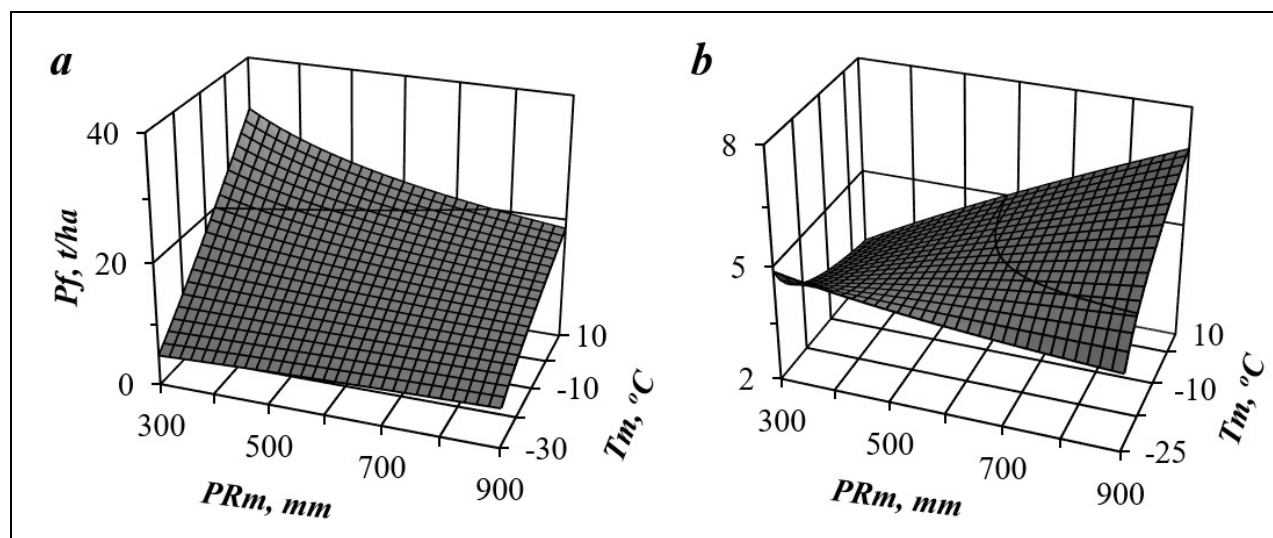


Fig. 2 Dependence of spruce (a) and oak (b) foliage biomass (P_f) of Eurasia upon the mean January temperature (T_m) and mean annual precipitation (PR_m).

We can see in Fig. 2 two patterns significant differed in changes in the spruce and oak foliage biomass in precipitation and winter temperature gradients in the form of two oppositely directed propellers. Similar oppositely directed 3D patterns were obtained earlier for larches and two-needled pines (Usoltsev et al., 2022).

Using the 3-D pattern in Fig. 2, we can find out, how much the foliage biomass will change with an air temperature deviation from the usual norm, for example, by 1°C and with a precipitation deviation from the usual norm, for example, by 100 mm per year. To do this, we take the first derivative of 3-D surfaces (Fig. 2), and not analytically, but graphically, i.e. we take off the biomass difference interval (Δ , %) corresponding to temperature interval 1°C and precipitation interval 100 mm directly from the graphs or from the corresponding tables, and get the answer in the form of 3-D surfaces divided into plus and minus areas that correspond to the increase or decrease in the foliage biomass for spruce (Fig. 2a) and for oak (Fig. 2b). It is assumed that climate change does not affect precipitation, which changes only geographically (by ecoregions), while as a result of the expected climate change the temperature increases by 1°C at different territorial (zonal) temperature levels, designated as -25Δ (-30Δ) ... $+10\Delta$. It is also assumed that climate change does not affect temperatures, which change only geographically (by ecoregions), while as a result of expected climate change, annual precipitation increases by 100 mm at various territorial precipitation levels, designated as 300Δ ... 900Δ .

As we can see on Fig. 3a, there is an increase in *Picea* foliage biomass (Δ , %) with a temperature increase of 1°C regardless of regional temperature and precipitation indicators (the location of the increment surface above the zero plane), and the largest percentage of increase occurs in cold regions ($T_m = -30^\circ\text{C}$) with insufficient moisture supply ($PR_m = 300\text{--}400$ mm), while *Quercus* leaf biomass in the same conditions, on the contrary, has the largest percentage of its decrease (the location of the increment surface below the zero plane) (Fig. 3b). In cold regions ($T_m = -30^\circ\text{C}$), as the transition from dry conditions ($PR_m = 300$ mm) to wet ones ($PR_m = 900$ mm), the percentage of foliage mass in spruce decreases, and in oak increases (Fig. 3a, b).

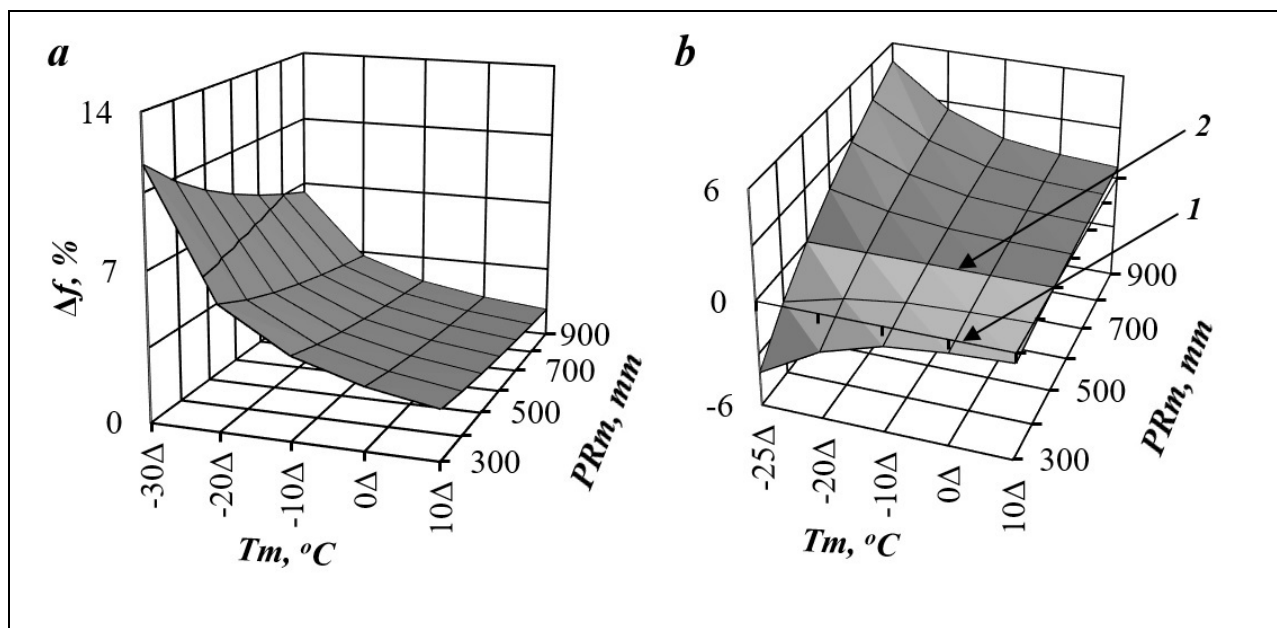


Fig. 3 Simulated changes in spruce (a) and oak (b) foliage biomass due to the assumed temperature increase of 1°C based on the derived model (for the stands aged 100 years). Here and further: the value 1 represents the plane corresponding to zero change of biomass at the expected temperature increase by 1°C; the value 2 represents the border between positive and negative changes in biomass ($\Delta, \%$) at the expected temperature increase by 1°C.

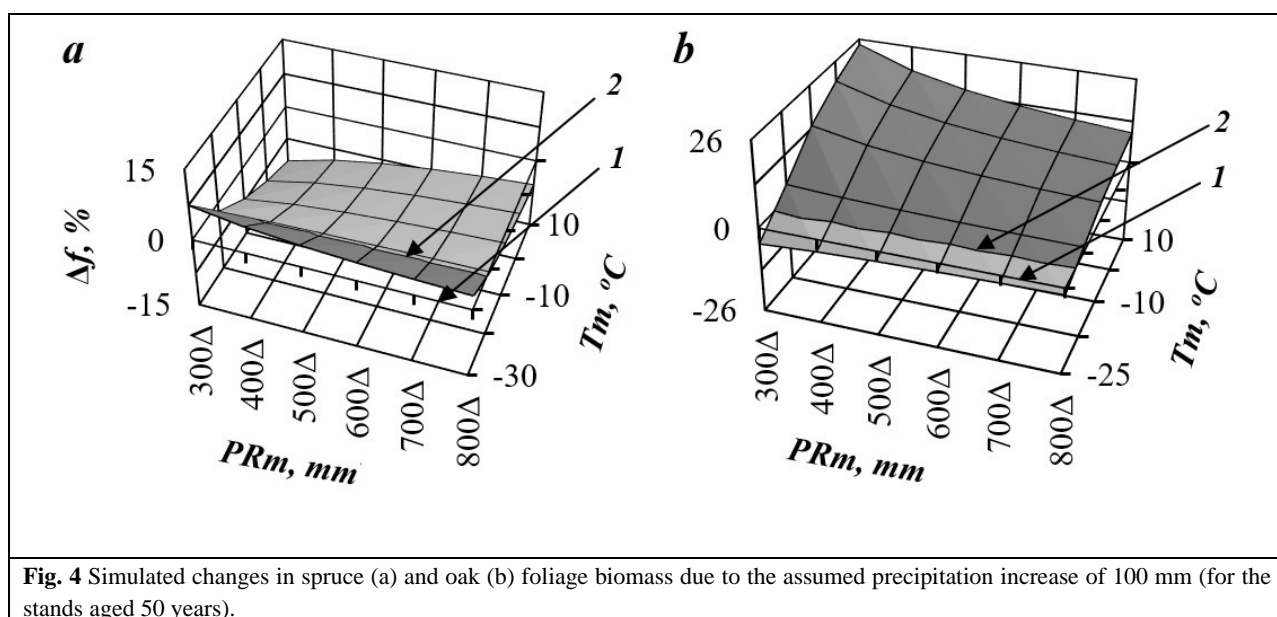


Fig. 4 Simulated changes in spruce (a) and oak (b) foliage biomass due to the assumed precipitation increase of 100 mm (for the stands aged 50 years).

If we assume an increase in precipitation by 100 mm at a constant winter temperature, we get exactly the opposite patterns for spruce and oak: in warm regions ($T_m = 10^\circ\text{C}$), spruce has a decrease in the percentage of foliage mass (Fig. 4a), and oak, on the contrary, its increase (Fig. 4b). As we move to cold regions ($T_m = -30^\circ\text{C}$), these patterns change to the opposite: the spruce negative percentage changes to positive, and the oak positive percentage changes to negative.

A next question: what can be the reason for the opposite trends in foliage biomass in spruce and oak, which are statistically reliable at the probability level equals 0.95 and even 0.99? Let us try to link the obtained counterintuitive patterns of changes in foliage biomass with previously published Trans-Eurasian trends in

foliage productivity as the ratio of annual growth in aboveground biomass to the biomass of the assimilation apparatus, that is known as foliage efficiency (FE) (Cannell et al., 1983; Shi and Cao, 1997). This index was analyzed in 5 natural zones (subarctic, northern temperate, southern temperate, subtropical, and subequatorial) with a stable Khromov's index of continentality (Khromov, 1957). It was found that in the direction from the northern moderate zonal belt to the subequatorial one, FE increases in spruce and decreases in oak in the same zonal range (Usoltsev et al., 2018). Another production index, foliage biomass, on the contrary, decreases in the direction from moderate to subequatorial zone in spruce, and increases in oak (Usoltsev, 2016). Opposite patterns of foliage biomass in spruce and oak were also observed in the continental climate gradient: a negative trend in spruce and a positive trend in oak (Usoltsev, 2016).

This phenomenon seems to be related to the fundamental difference between the winter physiology of evergreen and deciduous species, namely the ability of the former to assimilate atmospheric carbon dioxide and prolong the assimilation process beyond the vegetation period, which is usual for deciduous species (Jumelle, 1892; Ewart, 1896; Matthaei, 1902; Henrici, 1921; Iwanoff and Kossowitsch, 1929; Zacharowa, 1929; Ivanov and Orlova 1931; Printz, 1933; Cartellieri, 1935; Ålvik, 1939; Freeland, 1944; Zeller, 1951; Pisek and Rehner, 1958; Pisek, 1960; Lyr et al. 1974; Kramer and Kozlovsky 1983; Schaberg et al., 1995; Wieser, 1997; Smashevskiy, 2014).

In spruce trees in the southern direction, due to higher winter temperatures, the winter accumulation of assimilates increases as a result of excess of photosynthesis over respiration, which is associated with the autumn-winter litterfall of foliage (Zalesov et al., 1994), which apparently determines the trend of increasing FE in the direction from the northern temperate zone to the subequatorial one.

In oak trees, FE in the direction from the northern temperate to the subequatorial zone does not increase, as in spruce, but decreases, possibly due to an increase in the respiration losses during a shorter physiologically active period (due to leaf shedding) compared to spruce. In essence, everything depends on the ratio of photosynthesis and respiration in the spruce, especially during the period when the oak does not have photosynthesis, but only respiration.

If we assume the presence of a positive ratio of foliage biomass and FE, then in our case, the opposite FE trends of two species coincide with opposite trends in foliage biomass, but only in regions of insufficient moisture ($PRm = 300$ mm) (Fig. 2 a, b), and as we move to areas of sufficient moisture ($PRm = 900$ mm), this coincidence only in spruce, and in oak, this correspondence is replaced by a complete contradiction: the foliage mass increases during the transition from cold to warm regions, in contrast to the negative trend in the foliage efficiency.

The patterns of biomass amount change under assumed changed climatic conditions (Figs. 2-4) reflect long-term adaptive responses of forest stands to regional climatic conditions and do not take into account rapid trends of current environmental changes, which place serious constraints on the ability of forests to adapt to new climatic conditions (Givnish, 2002; Berner et al., 2013; Schaphoff et al., 2016; Spathelf et al., 2018; Vasseur et al., 2018; DeLeo et al., 2020; Denney and Anderson, 2020; Kannenberg et al., 2022). However, with a slow change in the environment, the effect of short-term stability of communities can be achieved (Baez et al., 2022). The law of limiting factors (Liebig, 1840) works well in stationary conditions. With a rapid change in limiting factors (for example, air temperature or precipitation), forest ecosystems are in a transitional state, in which some factors that were still not significant may come to the fore, and the end result may be determined by other limiting factors (Odum, 1971; Du et al., 2019).

The main pool of our harvest data on forest biomass in Eurasia was obtained since 1970s to 1990s, and the climate maps used cover the period of the late 1990s and early 2000s. Some discrepancy between the two time

periods may cause possible biases in the results obtained, but for such a small time difference in the initial data, the inclusion of compensatory mechanisms or phenological shifts in forest communities is unlikely (Anderegg et al., 2019; DeLeo et al., 2020). There is an uncertainty in assessing the impact of phenology on the biological productivity of stands, established for the cherry oak in the South of Russia: if the assessment of the biomass of oak stands did not reveal differences between the phenological varieties of oak, then the assessment of net primary production shows a 1.6-fold advantage of late-blooming variety over the early-blooming variety (Zhou, 1992).

Taking into account the stated methodological and conceptual uncertainties, the results presented in this study should be considered as preliminary ones. They can be modified if the biomass database will be enlarged by additional site-specific and stand-specific data. A full explanation of both the obtained regularities and counterintuitive results can be obtained after conducting detailed physiological studies, especially in relations between foliage biomass and foliage efficiency. Today, the most important question is whether the dryness of habitats will increase with climate warming. This determines the possible increase in carbon sink during climate warming, or, on the contrary, its decrease and the corresponding change to the source of carbon dioxide in the atmosphere (Tyrrell et al., 2012).

3 Conclusion

A comparison of the reaction of spruce and oak foliage biomass to changes in the mean January temperature by 1°C at constant precipitation and annual precipitation by 100 mm at constant temperature showed counterintuitive result. If low temperature regardless of regional precipitation is the limiting factor for spruce foliage biomass, then high temperature and insufficient precipitation are the limiting factors for oak.

When the temperature increases by 1°C, the most increase in spruce foliage biomass occurs in cold regions with insufficient moisture supply, while oak leaf biomass, on the contrary, has the largest percentage of its decrease. In cold regions, as the transition from dry conditions to wet ones, the percentage of foliage mass in spruce decreases, and in oak increases. When the precipitation increases by 100 mm, there is the opposite patterns for spruce and oak: in warm regions, spruce has a decrease in the percentage of foliage mass, and oak, on the contrary, its increase. In the cold regions, these patterns change to the opposite: the spruce negative percentage changes to positive, and the oak positive percentage changes to negative.

Our outputs represent an example of model sensitivity to changing climatic conditions. The development of such models for the main forest-forming species of Eurasia allow us to predict changes in the foliage productivity of the forest cover of Eurasia in relation to climate change.

References

- Ålvik G. 1939. Über Assimilation und Atmung einiger Holzgewächse im westnorwegischen Winter. *Meddelelser fra Vestlandets Forstlige Forsøksstation Bergen*, 6(4): 1-266
- Anderegg WRL, Anderegg LDL, Kerr KL, Trugman AT. 2019. Widespread drought - induced tree mortality at dry range edges indicates that climate stress exceeds species' compensating mechanisms. *Global Change Biology*, 25: 3793-3802
- Anderson KJ, Allen AP, Gilgooly JF, Brown JH. 2006. Temperature-dependence of biomass accumulation rates during secondary succession. *Ecology Letters*, 9: 673-682
- Baez S, Fadrique B, Feeley K, Homeier J. 2022. Changes in tree functional composition across topographic gradients and through time in a tropical montane forest. *PLoS ONE*, 17(4): e0263508
- Baruah G, Molau U, Bai Y, Alatalo JM. 2017. Community and species-specific responses of plant traits to 23

- years of experimental warming across subarctic tundra plant communities. *Scientific Reports*, 7: 2571
- Baskerville GL. 1972. Use of logarithmic regression in the estimation of plant biomass. *Canadian Journal of Forest Research*, 2: 49-53
- Berner LT, Beck PSA, Bunn AG, Goetz SJ. 2013. Plant response to climate change along the forest-tundra ecotone in northeastern Siberia. *Global Change Biology*, 19(11): 3449-3462
- Bijak S. 2010. Tree-ring chronology of silver fir and its dependence on climate of the Kaszubskie Lakeland (Northern Poland). *Geochronometria*, 35: 91-94
- Bobrov EG. 1978. *Forest-Forming Conifers of the USSR*. Nauka, Leningrad, USSR
- Bocharnikov MV. 2019. Bioclimate as a factor of the phytocenotic diversity's formation of the mountain taiga forests in the Stanovoy Highland. In: *Forest Ecosystems of Boreal Zone: Biodiversity, Bioeconomy, Ecological Risks*. Proceedings of the All-Russian Conference with Internat. 57-59, Krasnoyarsk, Russia
- Buckley LB, Kingsolver JG. 2012. Functional and phylogenetic approaches to forecasting species' responses to climate change. *Annual Review of Ecology, Evolution and Systematics*, 43: 205-226
- Cannell MGR., Sheppard LJ, Ford ED, Wilson RHF. 1983. Clonal differences in dry matter distribution, wood specific gravity and foliage "efficiency" in *Picea sitchensis* and *Pinus contorta*. *Sivae Genetica*, 32: 195-202
- Cartellieri E. 1935. Jahresgang von osmotischem Wert, Transpiration und Assimilation einiger *Ericaceen* der alpine Zwergstrauchheide und von *Pinus cembra*. *Jahrbuch für wissenschaftlich Botanik*, 82: 460-506
- Chernyshev VD. 1974. Pathways of physiological and energetic adaptations of conifers under extreme conditions. In: *Biologicheskoe problem Severa: Tezisy VI Simp. (Biological Problems of the North: Abstr. VI Symposium)*. Inst. Biol., Yakutsk. Fil. Sib. Otd. Akad., Yakutsk, Nauk SSSR, 5: 13-17
- DeLeo VL, Menge DNL, Hanks EM, et al. 2020. Effects of two centuries of global environmental variation on phenology and physiology of *Arabidopsis thaliana*. *Global Change Biology*, 26 (2): 523-538
- Denney DA, Anderson JT. 2020. Natural history collections document biological responses to climate change: A commentary on DeLeo et al. (2020), Effects of two centuries of global environmental variation on phenology and physiology of *Arabidopsis thaliana*. *Global Change Biology*, 26: 340-342
- Dolezal J, Jandova V, Macek M, Liancourt P. 2021. Contrasting biomass allocation responses across ontogeny and stress gradients reveal plant adaptations to drought and cold. *Functional Ecology*, 35(1): 32-42
- Dong Y, Liu Y. 2017. Response of Korean pine's functional traits to geography and climate. *PLoS ONE*, 12(9): e0184051
- Draper N, Smith H. 1966. *Applied Regression Analysis*. Wiley, New York, USA
- Du J, Watts JD, Jiang L, et al. 2019. Remote sensing of environmental changes in cold regions: Methods, achievements and challenges. *Remote Sensing*, 11: 1952
- Duan H, Huang G, Zhou S, Tissue DT. 2018. Dry mass production, allocation patterns and water use efficiency of two conifers with different water use strategies under elevated CO₂, warming and drought conditions. *European Journal of Forest Research*, 137(5): 605-618
- Dymond CC, Beukema S, Nitschke CR, et al. 2016. Carbon sequestration in managed temperate coniferous forests under climate change. *Biogeosciences*, 13: 1933-1947
- Eggers J, Lindner M, Zudin S, et al. 2008. Impact of changing wood demand, climate and land use on European forest resources and carbon stocks during the 21st century. *Global Change Biology*, 14: 2288-2303
- Emanuel WR, Shugart HH, Stevenson MP. 1985. Climate change and the broad scale distribution of terrestrial ecosystem complexes. *Climate Change*, 7: 29-43
- Ewart AJ. 1896. On assimilatory inhibition. *Journal of the Linnean Society of London, Botany*, 31(217):

364-461

- Fang O, Wang Y, Shao X. 2016. The effect of climate on the net primary productivity (NPP) of *Pinus koraiensis* in the Changbai Mountains over the past 50 years. *Trees*, 30: 281-294
- Felton A, Nilsson U, Sonesson J, et al. 2016. Replacing monocultures with mixed-species stands: Ecosystem service implications of two production forest alternatives in Sweden. *Ambio*, 45: 124-139
- Folland CK, Palmer TN, Parker DE. 2001. Climate change 2001: The scientific basis. In: Contribution of Working Group I to the Third Assessment Report of the Intergovernmental Panel on Climate Change (Houghton JT et al., eds). Cambridge University Press, Cambridge, UK
- Forrester DI, Tachauer IHH, Annighoefer P, et al. 2017. Generalized biomass and leaf area allometric equations for European tree species incorporating stand structure, tree age and climate. *Forest Ecology and Management*, 396: 160-175
- Frank D, Reichstein M, Bahn M, et al. 2015. Effects of climate extremes on the terrestrial carbon cycle: concepts, processes and potential future impacts. *Global Change Biology*, 21(8): 2861-2880
- Freeland RO. 1944. Apparent photosynthesis in some conifers during the winter. *Plant Physiology*, 19(2): 179-185
- Fu L, Sun W, Wang G. 2017. A climate-sensitive aboveground biomass model for three larch species in northeastern and northern China. *Trees*, 31: 557-573
- Givnish TJ. 2002. Adaptive significance of evergreen vs. deciduous leaves: solving the triple paradox. *Silva Fennica*, 36: 703-743
- Gower ST, Krankina O, Olson RJ, et al. 2001. Net primary production and carbon allocation patterns of boreal forest ecosystems. *Ecological Applications*, 11(5): 1395-1411
- Gower ST, Richards JH. 1990. Larches: Deciduous conifers in an evergreen world. *BioScience*, 40(11): 818-826
- Gray SB, Brady SM. 2016. Plant developmental responses to climate change. *Developmental Biology*, 419(1): 64-77
- Gu F, Yuandong Z, Mei H, et al. 2017. Effects of climate warming on net primary productivity in China during 1961-2010. *Ecology and Evolution*, 7(17): 6736-6746
- Gustafson EJ, Miranda BR, De Bruijn AMG, et al. 2017. Do rising temperatures always increase forest productivity? Interacting effects of temperature, precipitation, cloudiness and soil texture on tree species growth and competition. *Environmental Modelling & Software*, 97: 171-183
- Han SH, Kim S, Li G, et al. 2018. Effects of warming and precipitation manipulation on fine root dynamics of *Pinus densiflora* Sieb. Et Zucc. seedlings. *Forests*, 9: 14
- Henrici M. 1921. Zweigipfelige Assimilationskurven. Mitspezieller Berücksichtigung der Photosynthese von alpinenphanerogamen Schattenpflanzen und Flechten. *Verhandlungen der Naturforschenden Gesellschaft in Basel*, 32: 107-171
- Huang J, Tardif JC, Bergeron Y, et al. 2010. Radial growth response of four dominant boreal tree species to climate along a latitudinal gradient in the eastern Canadian boreal forest. *Global Change Biology*, 16: 711-731
- Hultén E. 1937. Outline of the History of Arctic and Boreal Biota During the Quaternary Period. J Cramer, New York, USA
- Huston MA, Wolverton S. 2009. The global distribution of net primary production: resolving the paradox. *Ecological Monographs*, 79(3): 343-377
- Ivanov LA, Orlova IM. 1931. To the question about the winter photosynthesis of our coniferous. *Zhurnal Russkogo Botanicheskogo Obshchestva*, 16(2-3): 139-157

- Iwanoff LA, Kossowitsch NL. 1929. Über die Arbeit des Assimilationsapparates verschiedener Baumarten. I. Die Kiefer (*Pinus silvestris*). *Planta*, 8: 427
- Jacobs MW, Cunia T. 1980. Use of dummy variables to harmonize tree biomass tables. *Canadian Journal of Forest Research*, 10: 483-490
- Jumelle MH. 1892. Recherches physiologiques sur les lichens. 2. Influence de basses températures sur l'assimilation. *Revue Générale de Botanique*, 4: 305-320
- Kannenbergs SA, Cabon A, Babst F, et al. 2022. Drought-induced decoupling between carbon uptake and tree growth impacts forest carbon turnover time. *Agricultural and Forest Meteorology*, 322: 108996
- Keeling HC, Phillips OL. 2007. The global relationship between forest productivity and biomass. *Global Ecology and Biogeography*, 16: 618-631
- Khan D, Din EU, Muneer MA, et al. 2019. Effect of temperature and precipitation on stem biomass and composition of white birch (*Betula platyphylla*) in Daxing'anling Mountain of Inner Mongolia, China. *Applied Ecology and Environmental Research*, 17(6):13945-13959
- Khitrovo AA. 1908. To the question of the fate of the oak forests of Central Russia. *Forest Journal (Lesnoy Zhurnal)*, 1: 42-58
- Khromov SP. 1957. K voprosu o kontinental'nostiklimata (To a problem of climate continentality). *Izvestiya Vsesoyuznogo Geograficheskogo Obshchestva*, 89(3): 221-225
- Korzhinsky SI. 1888. The northern border of the chernozem region of the Eastern strip of European Russia in botanical, geographical and soil terms. *Proceedings of the Society of natural scientists at Kazan University (Kazan) Part 1*, 18(5)
- Kramer PJ, Kozlovsky TT. 1967. *Physiology of Woody Plants*. Academic Press, New York, USA
- Kusmin SP. 1930. Water balance and drought-resistance of plants of the Apsheron peninsula in connection with peculiarities in the structure of their root systems. *Bulletin of Applied Botany, of Genetics and Plant-breeding*, 23(22): 393-421
- Laing J, Binyamin J. 2013. Climate change effect on winter temperature and precipitation of Yellowknife, Northwest Territories, Canada from 1943 to 2011. *American Journal of Climate Change*, 2: 275-283
- Lapenis A, Shvidenko A, Shepaschenko D, et al. 2005. Acclimation of Russian forests to recent changes in climate. *Global Change Biology*, 11: 2090-2102
- Lavrinenko DD. 1965. *Interaction of Tree Species In Different Types of Forest*. Lesnaya Promyshlennost Publishing, Moscow, USSR
- Levanič T, Poljanšek S, Toromani E. 2015. Early summer temperatures reconstructed from black pine (*Pinus nigra* Arnold) tree-ring widths from Albania. *Holocene*, 25(3): 469-481
- Lewis SL, Sonké B, Sunderland T, et al. 2013. Aboveground biomass and structure of 260 African tropical forests. *Philosophical Transactions of the Royal Society B Biological Sciences*, 368(1625): 20120295
- Liebig J. 1840. *Organic chemistry in its application to agriculture and physiology*. Braunschweig. In: German Text Archive. http://www.deutschestextarchiv.de/liebig_agricultur_1840. Accessed May 22, 2020
- Lieth H. 1974. Modeling the primary productivity of the world. *International Section for Ecology Bulletin*, 4: 11-20
- Lyr H, Polster H, Fiedler HJ. 1967. *Gehölzphysiologie*. VEB Gustav Fisher Verlag, Jena, Germany
- Matthaei GLC. 1902. The effect of temperature on carbon dioxide assimilation. *Annals of Botany*, 16: 591-593
- Mencuccini M, Manzoni S, Christoffersen B. 2019. Modelling water fluxes in plants: from tissues to biosphere. *New Phytologist*, 222(3): 1207-1222
- Monserud RA, Denissenko OV, Kolchugina TP, Tchebakova NM. 1996. Change in Siberian phytomass

- predicted for global warming. *Silva Fennica*, 30(2-3): article 5586
- Morley JW, Batt RD, Pinsky ML. 2017. Marine assemblages respond rapidly to winter climate variability. *Global Change Biology*, 23: 2590-2601.
- Morozov GF. 1949. *The Doctrine of The Forest* (7th edition). Goslesbumizdat Publishing, Moscow, Leningrad, USSR
- Ni J, Zhang X-S, Scurlock JMO. 2001. Synthesis and analysis of biomass and net primary productivity in Chinese forests. *Annals of Forest Science*, 58: 351-384
- Niu S, Wang S, Wang J, et al. 2020. Integrative ecology in the era of big data — From observation to prediction. *Science China Earth Sciences*, 63: 1-14
- Noce S, Caporaso L, Santini M. 2019. Climate change and geographic ranges: The implications for Russian forests. *Frontiers in Ecology and Evolution*, 7: 57
- Odum EP. 1971. *Fundamentals of Ecology* (The third Edition). W.B. Saunders Company, Philadelphia, London, Toronto
- Olsen N, Manda M. 2007. Will the Magnetic North Pole move to Siberia? *Transactions American Geophysical Union*, 88(29): 293-300
- Paquette A, Vayreda J, Coll L, et al. 2018. Climate change could negate positive tree diversity effects on forest productivity: A study across five climate types in Spain and Canada. *Ecosystems*, 21(5): 960-970
- Parmesan C, Hanley ME. 2015. Plants and climate change: complexities and surprises. *Annals of Botany*, 116: 849-864
- Pisek A. 1960. Immergrüne Pflanzen. *Handbuch der Pflanzenphysiologie*. Bd. V/2. , Springer, Berlin, Göttingen, Heidelberg
- Pisek A, Rehner G. 1958. Temperaturminima der Netto-Assimilation von mediterranen und nordisch-alpinen Immergrünen. *Berichte der Deutschen Botanischen Gesellschaft*, 71(4): 188-193
- Poudel BC, Sathre R, Gustavsson L, et al. 2011. Effects of climate change on biomass production and substitution in north-central Sweden. *Biomass and Bioenergy*, 35(10): 4340-4355
- Pretzsch H, Dieler J, Seifert T, Rötzer T. 2012. Climate effects on productivity and resource-use efficiency of Norway spruce (*Picea abies* [L.] Karst.) and European beech (*Fagus sylvatica* [L.]) in stands with different spatial mixing patterns. *Trees*, 26: 1343-1360
- Printz H. 1933. Granenes og furuens fysiologiske og geografiske udbredelse. *Nyt Magazin for Naturvidensk*, 73: 167-219
- Pucko C, Beckage B, Perkins T, Keeton WS. 2011. Species shifts in response to climate change: Individual or shared responses? *The Journal of the Torrey Botanical Society*, 138(2): 156-176
- Reich PB, Bolstad P. 2001. Productivity of evergreen and deciduous temperate forest. In: *Terrestrial Global Productivity* (Roy J, Saugier B, Mooney HA, eds). 245-277, Academic Press, San Diego, USA
- Reichstein M, Carvalhais N. 2019. Aspects of forest biomass in the earth system: Its role and major unknowns. *Surveys in Geophysics*, 40: 693-707
- Schaberg PG, Wilkinson RC, Shane JB, et al. 1995. Winter photosynthesis of red spruce from three Vermont seed sources. *Tree Physiology*, 15: 345-350
- Schaphoff S, Reyer CPO, Schepaschenko D, et al. 2016. Tamm Review: Observed and projected climate change impacts on Russia's forests and its carbon balance. *Forest Ecology and Management*, 361: 432-444
- Schulze D. 1982. Plant life forms and their carbon, water and nutrient relations. In: *Physiological Plant Ecology II* (Lange OL, Nobel PS, Osmond CB, Ziegler H, eds). 615-676, Springer-Verlag, Berlin, Germany
- Schulze ED, Mooney HA. 1994. *Biodiversity and Ecosystem Function*. Springer-Verlag, Berlin, Germany

- Shi K, Cao QV. 1997. Predicted leaf area growth and foliage efficiency of loblolly pine plantations. *Forest Ecology and Management*, 95(2): 109-115
- Shuman JK, Shugart HH. 2009. Evaluating the sensitivity of Eurasian forest biomass to climate change using a dynamic vegetation model. *Environmental Research Letters*, 4(4): 1-7
- Shuman JK, Shugart HH, O'Halloran TL. 2011. Sensitivity of Siberian larch forests to climate change. *Global Change Biology*, 17: 2370-2384
- Sippel S, Reichstein M, Ma X, et al. 2018. Drought, heat, and the carbon cycle: a Review. *Current Climate Change Reports*, 4(3): 266-286
- Smashevskiy ND. 2014. Ecology of photosynthesis. *Astrakhanskiy Vestnik Ekologicheskogoobrazovania (Astrakhan Bulletin of Environmental Education)*, 2(28): 165-180
- Spathelf P, Stanturf J, Kleine M, et al. 2018. Adaptive measures: Integrating adaptive forest management and forest landscape restoration. *Annals of Forest Science*, 75(2): 55
- Sperry JS, Venturasa MD, Todda HN, et al. 2019. The impact of rising CO₂ and acclimation on the response of US forests to global warming. *Proceedings of the National Academy of Sciences of the United States of America*, 116 (51): 25734-25744
- Stegen JC, Swenson NG, Enquist BJ, et al. 2011. Variation in aboveground forest biomass across broad climatic gradients. *Global Ecology and Biogeography*, 20: 744-754
- Strömngren M, Linder S. 2002) Effects of nutrition and soil warming on stem wood production of a boreal Norway spruce stand. *Global Change Biology*, 8: 1195-1204
- Sudachkova NE, Milyutina IL, Romanova LI. 2012. Biochemical Adaptation of Conifers To Stressful Conditions of Siberia. "GEO" Publishing, Novosibirsk, Russia
- Suvorova GG. 2009. Photosynthesis of Coniferous Trees Under The Siberian Conditions. "GEO" Publishing, Novosibirsk, Russia
- Suvorova GG, Shcherbatyuk AS, Yankova LS. 1999. Natural changes in photosynthetic productivity in conifers. In: *Methods for assessing the state and stability of forest ecosystems. Thesis of Reports*. 164-165, Institute of forest SB RAS, Krasnoyarsk, Russia
- Teets A, Fraver S, Weiskittel AR, Hollinger DY. 2018. Quantifying climate–growth relationships at the stand level in a mature mixed-species conifer forest. *Global Change Biology*, 24(8): 3587-3602
- Tolmachev AI. 1962. *Osnovyucheniyaobarealakh: Vvedenie v khorologiyurastenyi (Fundamentals of Plant Habitat Theory: Introduction to Plant Community Chorology)*. State University Publishing, Leningrad, USSR
- Toromani E, Bojaxhi F. 2010. Growth response of silver fir and Bosnian pine from Kosovo. *South-East European Forestry*, 1: 20-28
- Tyrrell ML, Ashton MS, Spalding D, Gentry B. 2012. Synthesis and conclusions. In: *Managing Forest Carbon in a Changing Climate* (Ashton MS, Tyrrell ML, Spalding D, Gentry B, eds). 377-387, Springer, Dordrecht, Heidelberg, London, New York
- Uoltsev VA. 1990. Mensuration of forest biomass: Modernization of standard base of forest inventory. In: *XIX World Congress Proceedings*. 79-92, IUFRO Division 4, Montreal, Canada
- Uoltsev VA. 2003. *Forest Biomass of Northern Eurasia: the limits of Productivity and Their Geography*. Ural Branch of Russian Academy of Sciences, Yekaterinburg, Russia
- Uoltsev VA. 2007. Some methodological and conceptual uncertainties in estimating the in-come component of the forest carbon cycle. *Russian Journal of Ecology*, 38(1): 1–10
- Uoltsev VA. 2016. *Biological Productivity of Forest-Forming Species In Eurasia's Climatic Gradients* (As

- Related To Supporting Decision-Making Processes In Forest Management). Ural State Forest Engineering University, Yekaterinburg, Russia
- Usoltsev VA. 2019. Forest Arabesques, or Sketches of Our Trees' Life (3rd Edition modified). Radomska Szkoła Wyższa w Radomiu, Radom, Poland
- Usoltsev VA. 2020. Forest biomass and primary production database for Eurasia: digital version. The third edition, enlarged. Ural State Forest Engineering University, Yekaterinburg, Russia
- Usoltsev VA, Shobairi SOR, Chasovskikh VP. 2018. Foliage productivity in Eurasian forests as related to climatic variables. *Discovery Nature*, 12: 91-101
- Usoltsev VA, Shakoor A, Zaib G, et al. 2022. Changes in foliage biomass of the genera *Larix* and *Pinus* along gradients of temperature and precipitation in Eurasia. *Pakistan Journal of Botany*, 54(4): 1-7
- Vasseur F, Exposito-Alonso M, Ayala-Garay OJ, et al. 2018. Adaptive diversification of growth allometry in the plant *Arabidopsis thaliana*. *Proceedings of the National Academy of Sciences of the United States of America*, 115: 3416-3421
- Vassiljev IM. 1931. Über den Wasserhaushalt von Pflanzen der Sandwüsteim südöstlichen Kara-Kum. *Planta*, 14: 225-309
- Wang Q, Zeng J, Leng S, et al. 2018. The effects of air temperature and precipitation on the net primary productivity in China during the early 21st century. *Frontiers of Earth Science*, 12(4): 818-833
- Wang WJ, He HS, Thompson III FR, et al. 2017. Changes in forest biomass and tree species distribution under climate change in the northeastern United States. *Landscape Ecology*, 32: 1399
- Whittaker RH, Likens GE, Lieth H. 1975. Scope and purpose of this volume. *Ecological studies: Analysis and synthesis*, 14: 3-4
- Wieser G. 1997. Carbon dioxide gas exchange of cembra pine (*Pinus cembra*) at the alpine timberline during winter. *Tree Physiology*, 17: 473-477
- Wilmking M, Juday GP, Barber VA, Zald HSJ. 2004. Recent climate warming forces contrasting growth responses of white spruce at treeline in Alaska through temperature thresholds. *Global Change Biology*, 10: 1724-1736
- World Weather Maps, 2007. <https://www.mapsofworld.com/referrals/weather/>
- Yarutkin IA. 1968. On the relationship between oak and spruce in the Middle Volga region. *Forest Science (Lesovedenie)*, 5: 24-31
- Ye Z, Yuan J, Damgaard J, et al. 2022. Climate warming since the holocene accelerates west–east communication for the Eurasian temperate water strider species *Aquarius paludum*. *Molecular Biology and Evolution*, 39(5): msac089
- Zacharowa TM. 1929. Über den Gasstoffwechsel der Nadelholzpflanzen im Winter. *Planta*, 8(1/2): 68-83
- Zalesov SV, Kryazhevskikh IA, Luganskiy NA. 1994. Component dynamics of the litter of bog pine forests and return of mineral elements with plant litter. *LesUrala i khozyaistvo v nikh*, 17: 102-111
- Zeller L, Liang J, Pretzsch H. 2018. Tree species richness enhances stand productivity while stand structure can have opposite effects, based on forest inventory data from Germany and the United States of America. *Forest Ecosystems*, 5: 4
- Zeller O. 1951. Über Assimilation und Atmung der Pflanzen im Winter bei tiefen Temperaturen. *Planta*, 39: 500-526
- Zhao D, Lynch TB, Westfall JA, Coulston JW. 2022. Additional biomass estimation alternatives: nonlinear two- and three-stage least squares and full information maximum likelihood for slash pine. *Canadian Journal of Forest Research*, 52: 1-14

- Zhou WN. 1992. Phytomass and features of nitrogen and ash elements cycling in the ecosystems of the cherry oak of various phenological varieties in the conditions of the Central Chernozem region. PhD Thesis. Voronezh Forest Engineering Institute, Voronezh, Russia
- Zubairov B, Heußner K-U, Schröder H. 2018. Searching for the best correlation between climate and tree rings in the Trans-Ili Alatau, Kazakhstan. *Dendrobiology*, 79: 119-130