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Constructing a synergistic regulation theory for enhancing natural enemy pest control through farmland vegetational diversity: From vegetation patches to functional networks

WenJun Zhang

School of Life Sciences, Sun Yat-sen University, Guangzhou 510275, China

E-mail: wjzhang@iaees.org, zhwj@mail.sysu.edu.cn

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Abstract

Enhancing vegetational diversity on farmland through practices such as weed belts, intercropping, trap cropping, and habitat establishment is widely recognized as a nature-based solution for suppressing pests and bolstering natural enemy communities. Nevertheless, the translation of this ecological principle into reliably effective and predictable pest management remains hindered by a persistent theoretical lag. Empirical outcomes oscillate between marked pest suppression and negligible effects, primarily because the underlying mechanisms are still conceptualized through static, univariate, and scale-insensitive hypotheses. This paper undertakes a comprehensive and critical synthesis of the existing knowledge base, rigorously deconstructing classical frameworks such as the enemies hypothesis, resource concentration hypothesis, and associational resistance concepts. Through a systematic chain of logical deductions anchored in network science, functional ecology, and non-equilibrium thermodynamics, I identify a fundamental conceptual flaw: the treatment of vegetation diversity as an undifferentiated bulk property rather than a spatially explicit, functionally complementary, and temporally orchestrated network. To rectify this, I propose an original theoretical construct, the Functional Vegetation Network Synergy Theory (FVNST). FVNST posits that sustainable pest suppression emerges only when three irreducible axes: functional module completeness, spatial network connectivity, and temporal resource continuity, simultaneously exceed critical thresholds, thereby triggering a synergistic, non-linear amplification of natural enemy efficacy. The theory is formalized through a suite of postulates and a mathematical model articulating the Natural enemy Support Potential and Pest Suppression Potential as a threshold-dependent function. This framework recasts the role of vegetation from a passive source of general biodiversity to an actively designed, multi-trophic ecological network infrastructure. The paper further delineates a five-step design methodology, introduces a quantitative Functional Vegetation Network Efficacy Index, and conceptually validates the theory through cross-system application scenarios. FVNST provides an engineering-ready, predictive core for conservation biological control, transforming it from a collection of empirical rules of thumb into a principled scientific discipline.

Keywords vegetational diversity; conservation biological control; functional vegetation network; synergistic regulation; trophic cascade; emergent properties; ecological thresholds.

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1 Introduction

1.1 Background and Problem Delineation

The sustainability of modern agriculture is fundamentally challenged by its reliance on synthetic chemical pesticides (Zhang, 2018b). The spectrum of negative externalities is extensive, encompassing the erosion of farmland biodiversity, contamination of soil and water resources, severe risks to human health, and the relentless evolution of pest resistance (Carson, 1962; Pimentel et al., 1992; Zhang, 2018b). This trajectory is ecologically and socially untenable, creating an urgent global imperative to transition towards pest management strategies that are both effective and inherently regenerative (Tilman et al., 2002; Ehler, 2006; Peshin et al., 2009; Tscharntke et al., 2012). Among the most promising alternatives is the deliberate enhancement of vegetational diversity within and around crop fields. Strategies such as maintaining weed belts, establishing vegetational shelters, integrating secondary plants, deploying trap crops, practicing intercropping, and constructing specific insectary habitats are archetypal nature-based solutions designed to amplify the ecological services provided by a diverse community of natural enemies, including predators, parasitoids, and entomopathogens (Heong et al., 1991; Naylor and Ehrlich, 1997; Landis et al., 2000; Altieri and Nicholls, 2004; Power, 2010; Meehan et al., 2011; Nicholls and Altieri, 2013; Paredes et al., 2015; Gurr et al., 2017; Zhang et al., 2014; van Lenteren et al., 2018; Zhang, 2025, 2026b, 2027a-b; Tiwari et al., 2024).

The fundamental premise is intuitively compelling and supported by a vast body of empirical evidence: diversified agroecosystems generally harbor a richer and more stable fauna of beneficial arthropods and exhibit lower herbivore pressure than their simplified counterparts (Shepard et al., 1987; Andow, 1991; Schoenly et al., 1998; Letourneau et al., 2011; Kremen and Miles, 2012; Zhang, 2007, 2011, 2012a-b, 2016a-b, 2018a, 2025; Singh, 2025). However, a deep and troubling schism separates this general principle from its reliable agronomic application. The practice of habitat management for pest control considerably predates a robust, predictive theory. For every documented success of a flower strip markedly increasing parasitism rates (Tschumi et al., 2015), a counter-example exists where enhanced diversity fails to translate into meaningful pest suppression, or worse, inadvertently benefits the pest (Straub et al., 2008; Tscharntke et al., 2005). This inconsistency reveals a central pain point: current design and implementation are overwhelmingly empirical and context-dependent, lacking a universal, mechanistic framework with genuine predictive power. Agricultural managers are left to navigate a landscape of ad-hoc recipes without a unifying theory to guide the strategic assembly of vegetation.

1.2 Research Aim

This study is conceived to directly address this theoretical vacuum. Its primary aim is to transcend the descriptive, inventory-based mindset of merely cataloging plant diversity metrics. Instead, it seeks to excavate the foundational, logical architecture that governs the translation of vegetational complexity into the reliable, emergent ecosystem function of pest suppression (Zhang, 2016a-b, 2018a). The core task is to construct, through a process of rigorous synthesis, logical deduction, and formal modeling, an original theoretical framework that can explain not just if, but how, when, and why a specific vegetational configuration will yield a stable and potent biocontrol outcome. This framework must be capable of answering the pragmatic questions of configuration: which functional elements are indispensable, in what spatial arrangement, and across which temporal window.

1.3 Innovation of This Study

This paper delivers several key conceptual and methodological innovations: (1) It reconceptualizes farmland vegetation not as a homogenous matrix of biodiversity, but as a dynamic ecological network composed of discrete, interacting functional modules (Zhang, 2011, 2012a-b, 2016b, 2018a, 2027a). (2) It proposes that the response of the natural enemy-pest system is not linear, but is governed by non-linear thresholds and the

principle of synergistic emergence (Zhang, 2016b); effective pest control is a higher-order system property that arises only when a specific set of component conditions are co-satisfied. (3) It establishes a triadic quantitative design logic that explicitly and simultaneously integrates spatial network topology, modularized trophic cascade dynamics, and temporal phenological coupling. By providing a formal grammar for this integration, the theory elevates conservation biocontrol from a craft towards an ecological engineering discipline.

2 Systematic Deconstruction of Existing Knowledge

2.1 Principal Modes of Vegetational Diversity and Their Analyzed Effects

The literature on agricultural habitat management can be systematically categorized based on the spatial geometry and functional intention of the vegetation relative to the crop matrix. This topological categorization is foundational to the subsequent functional classification.

2.1.1 Linear Boundary Types: Weedy Strips, Flower Belts, and Beetle Banks

Linear boundary features represent the most studied and implemented form of vegetational diversity (Zhang et al., 2014). Sown wildflower strips and conserved grassy margins are archetypal examples designed to provide pollen, nectar, and alternative prey for a wide array of natural enemies (Altieri et al., 1984; Altieri, 1996, 1999; Wäckers et al., 2004; Jonsson et al., 2008; Haaland et al., 2011; Balzan et al., 2016; Kleiman et al., 2021; Walston et al., 2024). The ecological dynamic is dominated by an edge effect and a spillover pattern, where natural enemy populations build up in the refuge habitat and subsequently disperse laterally into the adjacent crop to attack pest populations (IRRI, 1998; Qi, 2003; Bianchi et al., 2006; Blitzer et al., 2012; Zhang et al., 2015). Beetle banks, which are raised earth ridges sown with tussock-forming grasses, are specifically engineered to provide undisturbed overwintering sites for polyphagous predators like carabids and staphylinids, creating a persistent source pool for seasonal recolonization of fields (Landis et al., 2000; MacLeod et al., 2004). The functional efficacy of these linear elements is critically dependent on their spatial scale relative to the field, the permeability of the crop matrix to the dispersing agents, and the ratio of edge to interior area (Schellhorn et al., 2015; Martin et al., 2019).

2.1.2 Internal Mosaic Types: Intercropping, Trap Cropping, and In-field Plant Strips

Mosaic diversification integrates non-crop vegetation directly into the crop production area. Intercropping, the simultaneous cultivation of two or more crops, alters the apparency and microclimate of the habitat, directly engaging the resource concentration hypothesis by making host plants more difficult for specialist herbivores to locate (Root, 1973; Risch, 1981). Trap cropping is a more targeted strategy where a highly attractive secondary plant stand is deployed to lure pests away from the main cash crop, concentrating them for localized elimination by natural enemies or targeted insecticide application (Shelton and Badenes-Perez, 2006; Parolin et al., 2012). Furthermore, alternating strips of insectary plants within the crop, such as rows of buckwheat or alyssum, function as internal refueling stations for parasitoids and syrphids, dramatically shortening their foraging distance and increasing their residence time in the crop interior (Tylianakis et al., 2004; Gurr et al., 2017). The dynamics here are defined by intimate interspecific interactions involving mutualism, apparent competition, and direct competition for resources (Letourneau et al., 2011; Dassou and Tixier, 2016).

2.1.3 Island-Type Habitats: Beetle Banks, Shrub Islands, and Patches of Perennial Vegetation

Discrete, insular patches of non-crop vegetation, such as small woodlots, shrub thickets, or dedicated perennial refugia within an agricultural matrix, function as biodiversity reservoirs and source habitats (Tscharntke et al., 2005; Bianchi et al., 2006; Iamba and Teksep, 2021). Their ecological role is powerfully informed by the theory of island biogeography and metapopulation dynamics, where patch size and isolation are critical determinants of species colonization and extinction rates (MacArthur and Wilson, 1967; Hanski, 1999). A network of such patches can sustain a metapopulation of a specialist parasitoid that would otherwise be extinct

at a single, ephemeral field scale (Schellhorn et al., 2015; Rusch et al., 2016). These habitats provide critical ancillary functions, including secure overwintering sites, aestivation refuges from disturbance, and a source of spillover predators that recolonize the crop after disturbances like harvest or tillage (IRRI, 1998; Landis et al., 2000; Qi, 2003; Rand et al., 2006; Fig. 1).

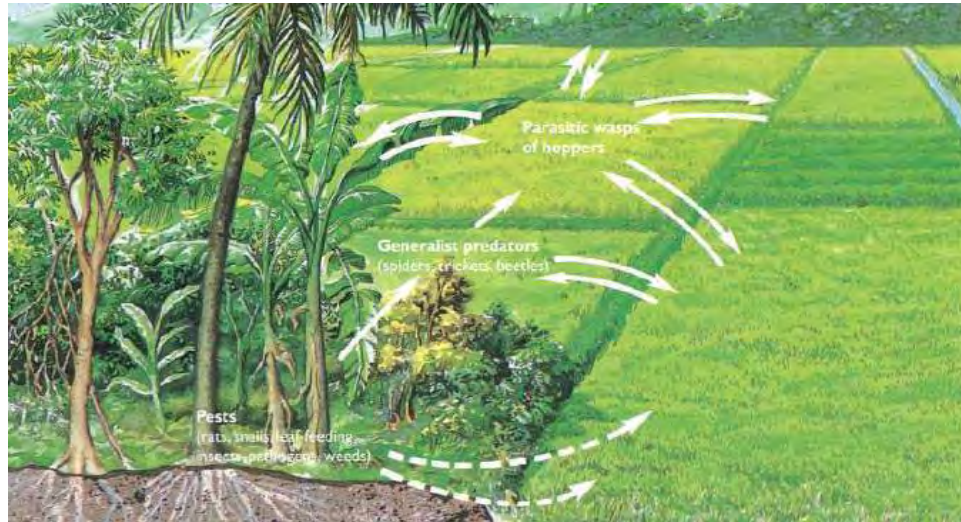


Fig. 1 Invertebrate diversity and natural pest control in rice paddies (IRRI, 1998). Showing interactions among vegetation habitats, rice, parasitic wasps, generalist predators (spiders, crickets, beetles), and pests (leaf-feeding insects, snails, etc.).

2.2 Classical Theoretical Hypotheses and Their Functional Logic

The interpretation of these vegetation-enhancement strategies has historically been governed by a set of interconnected ecological hypotheses. A deep analysis of their internal logic and limitations is essential to identify the theoretical bottlenecks.

2.2.1 The Enemies Hypothesis and Resource Concentration Hypothesis: A False Binary

Two dominant hypotheses have long framed the debate. The enemies hypothesis posits that vegetational diversity supports a greater abundance and effectiveness of natural enemies by providing alternative food sources, such as pollen and nectar for adults, shelter from hyperparasitoids, and a more stable microclimate, leading to stronger top-down control of herbivores (Root, 1973; Russell, 1989; Letourneau, 1987). In contrast, the resource concentration hypothesis argues that specialist herbivores are more likely to find, stay, and reproduce in dense, pure stands of their host plants, making monocultures more susceptible to pest outbreaks (Root, 1973; Andow, 1991). These hypotheses are often presented as a binary duality, yet this simplification is deeply limiting. The mere presence of alternative food sources does not automatically translate into enhanced biocontrol. The functional response is mediated by complex interactions such as intraguild predation, where one natural enemy preys upon another, potentially disrupting the trophic cascade and releasing the pest from control (Rosenheim et al., 1995; Finke and Denno, 2005). The two hypotheses operate concurrently and interact in a context-dependent manner, making the single-factor "diversity-good" or "monoculture-bad" heuristic insufficient for predictive design.

2.2.2 Associational Resistance and Susceptibility: The Unresolved Adjacency Effect

The concept of associational resistance extends the diversity dialogue to the neighborhood level, predicting that a plant's susceptibility to herbivore attack can be reduced when it is surrounded by non-host neighbors. This can occur through visual or olfactory camouflage, or by the neighboring plants harboring more effective natural enemies (Barbosa et al., 2009; Letourneau et al., 2011). Conversely, associational susceptibility posits

that neighboring plants can attract a generalist pest to the entire patch, increasing pressure on the target crop. The extensive meta-analyses (Meta-analysis: Zhang, 2024) and reviews on this topic reveal a high degree of contextual uncertainty; the outcome is highly sensitive to the specific combination of plant species, the feeding guild of the herbivore, and the spatial scale of observation (Andow, 1991; Barbosa et al., 2009). This unresolved contingency underscores a theoretical failure: current models lack the mechanistic links to predict the sign and magnitude of neighbor effects a priori.

2.2.3 Island Biogeography and Metapopulation Theory: A Static Spatial Template

Applying the equilibrium theory of island biogeography (MacArthur and Wilson, 1967) to agricultural landscapes conceptualizes habitat patches as islands and the crop matrix as a hostile ocean. This analogy has been crucial for understanding how patch size and isolation influence species richness (Zhang, 2026b), and metapopulation theory has provided a dynamic model for the persistence of species in fragmented habitats through a balance of colonization and extinction (Hanski, 1999; Tschamntke et al., 2005). The application to agricultural natural enemies, however, is often impoverished. The classic model treats the matrix as uniform and inhospitable and the species in a singular trophic level. It fails to account for the functional quality of the patch, the behavioral response of organisms to the inter-patch matrix structure, and the multi-trophic interactions that define the biocontrol function (Bianchi et al., 2006; Schellhorn et al., 2015; Zhang, 2016b). Consequently, while these theories explain patterns of species occurrence, they fall critically short of explaining the function of that diversity in delivering a biocontrol service.

2.3 Deep Bottlenecks and Identification of Theoretical Gaps

A systematic synthesis of the literature reveals a convergence of several profound, unresolved bottlenecks that constitute the theoretical vacuum.

The first bottleneck is the weak and often non-existent correlation between simple plant species diversity indices and effective pest control (Letourneau et al., 2011; Chaplin-Kramer et al., 2011). Existence does not guarantee function. A diverse assemblage of plants may be functionally redundant from a natural enemy's perspective, lacking the critical resource types necessary for its life-cycle completion. A perennial grass bank provides overwintering habitat but no floral resources, while a pure stand of a non-flowering annual provides nothing. Summing these up as a diversity index masks critical functional absences. The bottleneck is not one of quantity of diversity, but of the quality and complementarity of functional traits (Cadotte et al., 2011).

The second bottleneck is a fundamental mismatch of spatial scales. Most habitat management interventions are localized, implemented at the field or farm scale. In contrast, the key natural enemy species that provide the ecosystem service often operate on a scale of landscape mosaics, requiring a network of resources whose spatial arrangement determines metapopulation viability and spillover dynamics (Tschamntke et al., 2005; Kremen and Miles, 2012; Zhang, 2016b). A well-designed beetle bank in one field may fail if the surrounding landscape is so simplified that the regional pool of colonizing predators has been extirpated (Rusch et al., 2016). This scale mismatch is a principal source of the "context-dependency" that undermines predictable biocontrol.

The third bottleneck is the pervasive neglect of temporal dynamics. Natural enemy-pest-vegetation interactions are deeply embedded in time (Elsadany, 2012; Elhassanein, 2014; Rana, 2015; Shakil et al., 2015; Khan et al., 2019; Ranjith et al., 2019), governed by phenology, voltinism, and seasonal succession. The common practice of planting a single annual flowering strip that blooms for a brief four-week window creates a resource pulse followed by a long resource trough. Such designs induce a "pulse-crash" population dynamic in the natural enemy community, failing to provide the continuous, year-round resource continuum required for sustained population maintenance (Landis et al., 2000; Wäckers et al., 2007). The phenological synchrony and temporal match-mismatch between critical resource supply and natural enemy demand are rarely

incorporated as a design parameter.

The final bottleneck is the persistent black box of multi-trophic interactions. Vegetation does not simply add a neutral resource layer; it fundamentally restructures the entire interaction network. The introduction of alternative prey or floral resources can either strengthen top-down control through apparent competition or weaken it by diverting predators from the target pest (Straub et al., 2008). Similarly, higher trophic levels like hyperparasitoids can be inadvertently promoted by the same floral resources intended for the primary parasitoids, thereby decoupling the trophic cascade (Araj et al., 2009; Jonsson et al., 2008; Jiang and Zhang, 2015). The mechanisms by which vegetational patches reconfigure the robustness and trophic efficiency of the entire food web remain largely undeconstructed and absent from management models.

3 Rigorous Logical Deduction: The Deductive Chain Leading to an Original Theory

To break through these accumulated theoretical bottlenecks, a systematic logical deduction is required, moving from a redefinition of fundamental categories to the formal derivation of emergent properties. This chain of reasoning establishes the logical necessity for the new theory.

3.1 From Vegetational Diversity to A Set of Functional Modules: A Functional Ontological Reconstruction

The foundational logical error in the current paradigm is a categorical one: the treatment of vegetation as an aggregated, passive input variable quantified by species count or a Shannon index. Natural enemies do not respond to diversity per se; they respond to the spatio-temporal continuity of a discrete set of indispensable resources and conditions. A parasitoid's life-cycle completion requires a specific sequence of resources: nectar for adult maintenance (Wäckers and van Rijn, 2012), a host for oviposition, alternative hosts during crop absence (Settle et al., 1996), a prey for its larval stage, and a secure overwintering refuge (Landis et al., 2000). The absence of any single, critical resource creates a bottleneck that leads to local extinction (Zhang, 2026c), independent of the richness of non-critical plants. This necessitates a functional ontology. The universe of vegetational diversity must be partitioned into irreducible, function-specific modules. Based on their primary service to the natural enemy community, four archetypal modules can be deduced. The Nourishment Module provides pollen, nectar, and alternative prey or hosts, serving as the metabolic engine for natural enemy populations (Gurr et al., 2017). The Shelter Module provides physical habitat for overwintering, aestivation, and protection from disturbance or hyperparasitism; this includes beetle banks, tussock grasses, and perennial hedgerows (Landis et al., 2000; MacLeod et al., 2004; Zhang et al., 2014; Zhang, 2025, 2026d). The Connectivity Module functions as a dispersal corridor or stepping-stone through the hostile crop matrix, reducing the effective distance between resource patches and facilitating spillover and metapopulation rescue effects (Schellhorn et al., 2015; Balzan et al., 2016). The Defense Module includes trap crops and repellent plants that directly manipulate pest behavior, concentrating them for annihilation or diverting them from the main crop (Shelton and Badenes-Perez, 2006; Parolin et al., 2012). This ontological shift from a bulk property of diversity to a minimal set of functionally complementary modules is the first, non-negotiable step in any engineering design.

3.2 Source-Sink Mosaics and a Critical Functional Connectivity Threshold: Spatial Ecological Network Reasoning

The second step in the logical chain addresses the spatial configuration of these functional modules. A set of isolated, high-quality modules, no matter how complete, does not constitute a functional system. The persistence and efficacy of a natural enemy population across a dynamic agricultural landscape is a problem of spatial flows. Drawing on the principles of metapopulation theory (Hanski, 1999) and landscape ecology (Taylor et al., 1993), the landscape can be modeled as a graph (Zhang, 2012a, 2016b, 2018a), where nodes are

patches of functional vegetation modules and edges represent the permeability of the intervening crop or non-crop matrix to the movement of the biological control agent. The source-sink dynamics are critical. A large, high-quality shelter module may act as a source population, exporting a surplus of natural enemies to surrounding agricultural sinks, where reproduction may be insufficient to offset mortality from disturbance (Dunning et al., 1992; Bianchi et al., 2006). The persistence of the metapopulation, and hence its ability to provide consistent pest control, is not a linear function of total habitat area. Graph theory and percolation theory indicate the existence of a critical threshold in network connectivity (Taylor et al., 1993; Zhang, 2016b, 2016d, 2018a), a critical functional connectivity denoted as Φ_c . When the structural and functional connectivity of the landscape graph falls below this percolation threshold, the network becomes fragmented into isolated components that are too small and too far apart to sustain viable populations; the system is in a state of chronic demographic failure, characterized by repeated local extinctions without recolonization (Schellhorn et al., 2015; Holland et al., 2016). As connectivity is increased, for instance by the strategic placement of connectivity modules or softening the matrix, the network transitions across a phase change. Above Φ_c , a single, large, spanning cluster emerges, allowing for the free movement of organisms and a regional rescue effect that dramatically reduces the probability of global metapopulation collapse (Zhang, 2012a, 2016b, 2018a). The logical deduction is thus precise: the transition from a merely temporary, immigration-dependent presence of natural enemies to a persistent, self-sustaining network is a threshold phenomenon, triggered only when the functional connectivity Φ of the designed vegetation network surpasses the critical value Φ_c .

3.3 From Linear Chains to Modular Trophic Networks: Logical Dimension Elevation of the Trophic Cascade

The third logical deduction transcends the classic, three-level linear trophic cascade (plant-herbivore-predator). The introduction of functional vegetation modules does not just add a fourth, basal trophic level; it fundamentally restructures the interaction topology of the entire community into a modular network. Classical food webs in simple monocultures tend to be dominated by few strong links, making them vulnerable to invasive pest outbreaks and exhibiting chaotic dynamics that preclude stable, low-density pest regulation (McCann, 2000; Neutel et al., 2002; Zhang, 2011). The nourishment and shelter modules function by providing alternative prey and refuges that create weak-to-moderate strength interaction links, particularly in the form of apparent competition between the pest and alternative prey. This structural change is not random; it introduces modularity into the trophic network. A modular network structure is characterized by a high density of interactions within certain groups (modules) and a lower density of interactions between groups (May, 1974; Thebault and Fontaine, 2010; Zhang, 2012a, 2016b-c, 2018a). Specifically, the functional modules can create strong intra-module coupling, such as a strong predator-alternative prey interaction within a beetle bank, that generates a stable, season-long source of generalist predators. These predators then exert a more diffuse, weakly coupled interaction on the target pest in the crop module. This modular architecture has a robust stabilizing effect. It allows perturbations, like a sudden pest outbreak, to be absorbed and dampened within its module without cascading uncontrollably and destabilizing the entire network (Krause et al., 2003; Stouffer and Bascompte, 2011). This logic leads to the formulation of a new proposition, the Modular Trophic Robustness Hypothesis: the efficacy and stability of pest suppression is maximized when the functional vegetation network is configured to induce a modular trophic network architecture, characterized by strong intra-modular and weak inter-modular energy flow coupling, which channels a generalist predator's functional response efficiently and stably towards the pest.

3.4 The Inevitability of Temporal Dynamics: Niche Temporal Relay and a Successional Continuum of Phenological Functions

The logic of spatial and trophic network design collapses without a temporal dimension. The agricultural year presents a fluctuating abiotic and biotic landscape. A snapshot design that provides abundant nectar for a single month creates a resource pulse that inevitably generates a pulse-crash population cycle in its dependent consumers (Rauwald and Ives, 2001; Schellhorn et al., 2015). Following this logic, such a design is not just suboptimal, it is a design for failure, locking the natural enemy community into an unstable dynamic that oscillates between a high, non-sustainable peak and a deep trough coinciding with the next pest outbreak. Sustained, high-density pest suppression requires that the vegetational infrastructure generates a continuous, unbroken spectrum of functional resources. This can only be achieved by an orchestrated temporal relay of different plant species and functional modules, timed to the phenology of the key natural enemy species. This requires a shift from selecting plants based on species identity to selecting a community composition that maximizes the temporal continuity index T of the functional services, particularly the provision of nectar and alternative hosts from early spring through late autumn. This temporal continuity creates a persistent, elevated baseline population of natural enemies that can immediately respond in a density-dependent manner to incipient pest outbreaks, instead of having to rebuild from zero (Crawley, 1992). This deduces a necessity: any functional vegetation network that fails to achieve a high T , suffering a resource trough, will inevitably experience a period of natural enemy scarcity, creating a window of vulnerability and a system that toggles between biocontrol success and failure.

3.5 Convergence of Core Corollaries: The Conditional Spectrum for Emergent Pest Suppression

The three independent logical chains converge on a single, profound conclusion. The reliable, sustainable suppression of pests is not a property of any single component of the system, be it a high diversity index, a large habitat area, or a specific plant species. It is an emergent system-level property. The biocontrol service emerges only when a specific set of non-negotiable conditions co-occur in a synergistic configuration. The complete condition spectrum requires the co-satisfaction of an irreducible functional module integrity M , a network connectivity surpassing a critical threshold $\Phi > \Phi_c$, and a temporal continuity T approaching unity. The failure in any one axis causes a systemic failure, a catastrophic collapse of the biocontrol function, regardless of the performance in the other two dimensions. This is the nature of synergistic regulation: the whole is not merely more than, but qualitatively different from, the sum of its parts. The logic therefore mandates a new theoretical framework that can formalize this conditional emergence.

4 Construction of the Original Theory: The Functional Vegetation Network Synergy Theory (FVNST)

Driven by the deductive logic established above, this section formally presents an original theoretical framework: the Functional Vegetation Network Synergy Theory (FVNST; Fig. 2). This theory offers a mechanistic, non-linear, and predictive account of how designed vegetational complexity suppresses pests.

4.1 The Theory and Core Postulates

FVNST is built upon a set of four irreducible and interdependent core postulates that define the necessary and sufficient condition set for the emergence of robust biocontrol. These postulates transform the metaphorical language of "diversity" into a set of testable, engineering parameters.

Postulate 1: Irreducible Functional Module Completeness. A stable and effective biocontrol system must be composed of a minimal, functionally complementary set of modules. The core triad consists of a Nourishment Module (M_N), a Shelter Module (M_S), and a Connectivity Module (M_C). The absence of any one of these module types creates a fatal bottleneck that will lead to the functional collapse of the natural enemy community across the system, a regime shift back to a pest-dominated, low-control state (Altieri and Nicholls, 2004; Wäckers and van Rijn, 2012). The system's functional completeness index M cannot be assessed by the size of one module but by the presence and minimal viability of all three.

Postulate 2: Spatial Network Threshold Trigger. The persistent, self-sustaining establishment of a natural enemy metapopulation and its associated pest control service is a non-linear, threshold-triggered phenomenon. It is only activated when the functional connectivity (Φ) of the landscape network, representing the effective permeability between the triadic functional modules, exceeds a system-specific critical threshold (Φ_c). Below this percolation point, the regional population is demographically unsustainable, collapsing to extinction. Above it, the metapopulation enters a high-viability, self-reinforcing phase (Hanski and Ovaskainen, 2000; Schellhorn et al., 2015). Pest suppression potential is thus a sigmoidal, not linear, function of connectivity.

Postulate 3: Modular Trophic Network Stabilization. The functional network achieves maximum and most robust pest suppression when its architecture induces a modular trophic cascade structure within the recipient food web. This structure is defined by high interaction strength within functional modules (e.g., a strong loop between alternative prey and a generalist predator in a shelter module) and weaker, more diffusive interaction strength between the generalist predator module and the target pest in the crop module (Neutel et al., 2002; Stouffer and Bascompte, 2011). This topology maximizes the channeling of energy from generalist predators toward the pest during a pest outbreak (apparent competition) while simultaneously preventing the destabilizing oscillations (paradox of enrichment) that characterize linear, strong-coupling food chains.

Postulate 4: Phenological Resource Continuum. The functional output of the vegetation network must be engineered as an unbroken temporal continuum of critical resources, particularly alternative prey and nectar. A phenological break or resource trough, where a key resource is absent, constitutes a systemic fracture. This fracture causes the natural enemy population to enter a deterministic decline, reverting the system to a pulse-crash dynamic and stripping it of the baseline high-density predator community needed for immediate pest suppression. The system's temporal continuity index T must be maintained above a high threshold ($T > 0.8$) for the network to be a permanent, regulating structure rather than a temporary, perturbation-dependent subsidy (Landis et al., 2000; Wäckers et al., 2007).

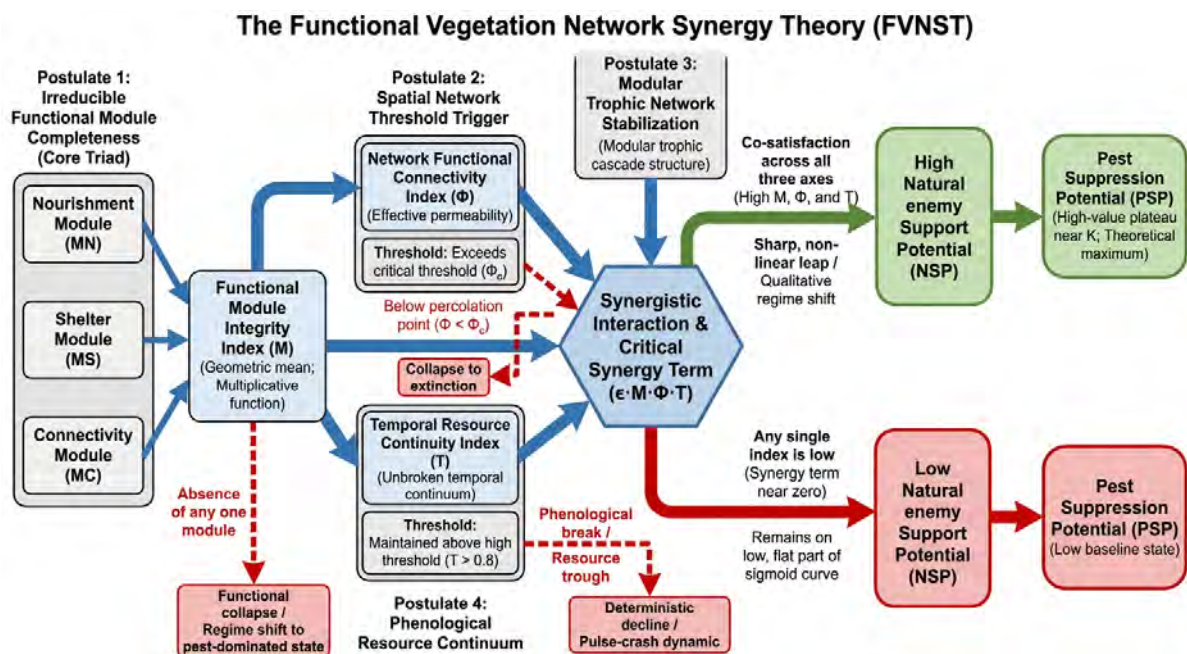


Fig. 2 Systematic diagram of The Functional Vegetation Network Synergy Theory (FVNST).

4.2 Theoretical Model Skeleton

These postulates can be synthesized into a formal mathematical function that captures the synergistic and non-linear nature of the biocontrol service. I define a key latent variable, the Natural enemy Support Potential (*NSP*), and its emergent outcome, the Pest Suppression Potential (*PSP*).

Let M be the Functional Module Integrity Index, a composite score ranging from 0 to 1, calculated as a multiplicative function of the presence and quality of the Nourishment, Shelter, and Connectivity modules. Its simplest form is the geometric mean of the quality scores of the three modules, which penalizes the absence of any module with a zero score, reflecting the logic of Postulate 1.

Let Φ be the Network Functional Connectivity Index, a graph-theoretic metric on the landscape scale (0 to 1), representing the probability of successful dispersal between functional modules. Let T be the Temporal Resource Continuity Index, measured as the proportion of the total period of natural enemy activity during which a critical resource (e.g., nectar, alternative hosts) is available (0 to 1).

The *NSP* is then a synergistic function of these three parameters. The core of the theory is that *PSP* does not increase additively with these components, but through a thresholded, synergistic interaction. A robust formalization of this is a multivariate logistic or Hill function:

$$PSP = K / [1 + \exp(-(\alpha(M - M_{1/2}) + \beta(\Phi - \Phi_{1/2}) + \gamma(T - T_{1/2}) + \epsilon \cdot M \cdot \Phi \cdot T))]$$

Where

K is the theoretical maximum pest suppression potential, an asymptote.

α, β, γ are scaling coefficients that determine the steepness of the transition along each axis.

$M_{1/2}, \Phi_{1/2}, T_{1/2}$ are the half-saturation constants, representing the points at which each individual variable has reached half its maximum potential effect in isolation.

$\epsilon \cdot M \cdot \Phi \cdot T$ is the critical synergy term. This term mathematically encodes the central FVNST prediction: when all three indices are simultaneously high, the argument of the exponential function becomes a large negative number because of this synergistic product term, causing the *PSP* to undergo a sharp, non-linear leap from a low baseline to a high-value plateau near K . If any single index is low, the synergy term is near zero, and the *PSP* will remain on the low, flat part of the sigmoid curve, regardless of how high the other two indices are.

This model formalizes the emergent threshold; high performance in a single axis yields marginal returns, while co-satisfaction across all three axes triggers a synergistic, qualitative regime shift into a high-*PSP* state.

4.3 The Theory's Original Contribution and Transcendence

FVNST constitutes a fundamental advance beyond the classical hypotheses. The enemies hypothesis (Root, 1973) and the associational resistance concept (Barbosa et al., 2009) are descriptive correlates; they explain patterns without specifying the engineering conditions for their reliable realization. FVNST provides these conditions as a set of non-negotiable thresholds. It transforms the passive, "more is better" diversity heuristic into an active, precise functional configuration engineering principle. Moreover, FVNST is the first framework to unify the principles of metapopulation rescue effects, modular food web stability, and phenological coupling into a single, mechanistic threshold model. It explicitly resolves the problem of context dependency by locating the source of variation not in an undefinable environment, but in the state of three measurable system axes relative to their critical thresholds. This provides a theoretical kernel for transforming agricultural ecological measures from an empirical craft into a designable, predictable, and diagnosable systems engineering discipline.

5 Methodology and Application Logic Based on the New Theory

Translating FVNST from a conceptual framework into actionable practice requires a structured design methodology and a set of quantitative decision-support tools. This section outlines a protocol for synthesizing a FVNST-compliant functional vegetation network.

5.1 A Five-Step Method for Functional Vegetation Network Design

The design process follows the logical chain of the theory itself, iteratively constructing and validating the network.

Step 1: System Diagnosis and Functional Blueprinting. The first step is to deconstruct the focal agroecosystem. A functional blueprint is created by identifying the key pest species, the critical natural enemy complex, and their specific, limiting resource bottlenecks throughout the year. This involves mapping the life-cycle vulnerabilities: when does a key parasitoid need nectar, where does a key carabid overwinter, and what alternative prey sustains a generalist mirid before the pest arrives (Settle et al., 1996; Gurr et al., 2017)? This diagnostic phase defines the functional specifications for the modules, the species-level requirements that must be met by the vegetation.

Step 2: Building the Functional Module Portfolio. Based on the blueprint, candidate plant species and structural elements are selected to constitute the three irreducible modules. The Nourishment Module is assembled with a portfolio of plants whose functional traits (e.g., short-corolla flowers for parasitoids, extrafloral nectaries) match the specifications (Wäckers, 2004). The Shelter Module is designed with specific architectures, such as perennial grass tussocks for overwintering carabids (MacLeod et al., 2004). The Connectivity Module is conceptualized as linear elements, such as grassy field margins or contour strips, that facilitate movement (IRRI, 1998; Zhang et al., 2014).

Step 3: Spatial Network Optimization and Connectivity Verification. The candidate modules must now be spatially arranged to ensure the connectivity threshold ($\Phi > \Phi_c$) is surpassed. The agricultural landscape is rendered as a spatial graph. Using GIS and least-cost path analysis (Zhang, 2017, 2018a), the functional resistance of the crop matrix to the movement of the indicator natural enemy species is mapped. Graph-theoretic metrics such as the integral index of connectivity or the probability of connectivity (Zhang, 2012a-b, 2016b, 2016d, 2018a) are calculated for the proposed spatial configuration (Pascual-Hortal and Saura, 2006). The design is iteratively optimized, adding stepping-stone habitats or widening corridors, until the modelled landscape connectivity indices surpass the predetermined critical threshold for the focal organism, guaranteeing network percolation.

Step 4: Phenological Matching and Temporal Continuum Design. The selected plant community is subject to a phenological audit. A monthly resource-availability calendar for each functional module is constructed. Gaps in the nectar supply or alternative prey continuity are identified. The design is refined by substituting or adding plant species with complementary phenologies, such as early-season Brassicaceae followed by Apiaceae and late-season Asteraceae, to engineer a high temporal continuity index $T > 0.85$, ensuring the permanent presence of the functional resource (Landis et al., 2000; Wäckers and van Rijn, 2012).

Step 5: Modular Network Resilience Validation. The finalized design is stress-tested *in silico*. The stability of the trophic network (Carpenter et al., 2001; Zhang, 2016a-b, 2018a) and the modeled *PSP* function are assessed under simulated perturbation scenarios. A typical test involves removing a component, such as the failure of the Nourishment Module due to a drought, and observing the predicted impact on *PSP*. A robust FVNST design will exhibit a degree of functional redundancy, where the Shelter Module's alternative prey service partially compensates, preventing a catastrophic decline of *PSP* to zero and demonstrating network resilience (Cumming, 2011; Zhang, 2016b).

5.2 Quantitative Evaluation Indices and Basis

For rapid field-scale assessment and monitoring, I propose a composite metric: the Functional Vegetation Network Efficacy Index (*FVNEI*). This index is a practical operationalization of the *PSP* model. It is calculated as a triple product of the normalized, complementary indices derived from the core postulates:

$$FVNEI = M_{norm} \cdot \Phi_{norm} \cdot T_{norm}$$

Where each sub-index is a unitless score from 0 to 1. For a rapid expert-based assessment, *M* can be scored based on a checklist verifying the presence and quality of a Nourishment, a Shelter, and a Connectivity module. Φ can be approximated by a simplified metric of edge density of the connectivity corridors and the permeability of the crop matrix. *T* is scored by assessing the proportion of the season with active floral resources. The *FVNEI* provides a single, communicable number that reflects the synergistic potential of the system. An *FVNEI* value below a calibrated threshold indicates a system stuck in a low-control basin of attraction, requiring a design intervention to boost one or more of the deficient axes.

5.3 Conceptual Verification Reasoning for Multiple Systems

The explanatory and predictive power of *FVNST* can be demonstrated through the conceptual re-interpretation of well-documented agroecological systems. In an irrigated rice ecosystem, a classic biological control success involves the manipulation of non-crop vegetation. The practice of maintaining vegetated bunds rich in Poaceae and Cyperaceae can be functionally deconstructed (Zhang et al., 2014) Shelter Module for a community of generalist predators, including spiders and mirids. By providing a continuous source of alternative prey, detritivores, and plant hopper species that do not attack rice, the bunds function as a potent Nourishment Module (Settle et al., 1996; Way and Heong, 1994). The linear network of bunds directly serves as the Connectivity Module, allowing rapid recolonization of rice paddies after pesticide application. In *FVNST* terms, this traditional configuration often achieves a high *M* and a high Φ at the field complex level, triggering the non-linear transition to stable pest suppression observed in many farmer-practice systems that avoid early-season broad-spectrum insecticides. The failure of such a system can be diagnosed by a temporal break in the *T*-index caused by herbicide application that destroys the Nourishment Module, proving the theory's diagnostic value.

A second application is an organic orchard system. The design integrates a ground-cover of a functionally diverse perennial legume mixture (Nourishment and Shelter Module), a network of insectary hedgerows (Connectivity and additional Nourishment Module), and a barrier of trap-crop plants (Defense Module). The ground cover provides alternative prey like aphids and mites that sustain a resident population of generalist predators like earwigs and anthocorids (Dib et al., 2016). The hedgerows provide a structurally complex corridor network linking the orchard floor to the tree canopy. *FVNST* predicts that such a system achieves pest suppression only because the spatial arrangement of these three module types triggers a strong synergistic effect, maintaining predators at high densities in the canopy early in the season, whereas an orchard with a ground cover but no connected, perennially flower-rich hedgerow would have a sub-critical Φ , limiting the predator's ability to move to the canopy and failing to achieve control.

6 Discussion

6.1 Compatibility and Upgrade of Classical Hypotheses

The Functional Vegetation Network Synergy Theory does not invalidate its predecessors; it encapsulates, refines, and upgrades them by placing them within a broader conditional architecture. The enemies hypothesis (Root, 1973) is correct in its assertion that diversity can support more natural enemies, but *FVNST* clarifies that it is specifically a functionally complementary, networked diversity that does so. The resource

concentration hypothesis is not negated but is recognized as one mechanism operating within the Defense Module, where a trap crop concentrates the pest. The associational resistance phenomenon (Barbosa et al., 2009) is deconstructed into its network components: an olfactory or physical disruption that is a form of spatial connectivity management. FVNST provides the precise engineering language that these classic hypotheses lack, specifying the necessary and sufficient conditions (M , Φ , T) under which their predicted effects will actually manifest in an agronomically meaningful way. It explains the variance in the literature: many experiments that claimed to test the enemies hypothesis actually involved systems with sub-critical M , Φ , or T , and thus, according to FVNST, were never predicted to yield a positive result in the first place.

6.2 Applicability Boundaries and Potential Limitations of the Theory

FVNST, as a spatial and ecological engineering theory, has defined boundaries. Its implementation in extremely simplified and structurally homogenized landscapes, such as vast cereal monocultures on the steppe, poses severe challenges. The source pool of natural enemies may be so regionally depleted that the connectivity threshold cannot be met without a massive, economically prohibitive landscape redesign (Tscharntke et al., 2005; Rusch et al., 2016). A second limitation is the availability of a species pool of functional plants that can fulfill all module requirements across different biogeoclimatic zones. Transferring the specific plant species composition from one region to another is invalid; the functional trait specifications must be matched to the local flora, a process requiring deep local natural history knowledge. Furthermore, the socio-economic constraints of large-scale implementation are significant. There is a high initial knowledge and capital cost to designing and installing a functional network, and the benefits of pest suppression may take several seasons to fully accrue as the natural enemy metapopulation stabilizes, creating a discount-rate barrier for farmers.

6.3 Implications for the Science of Sustainable Agriculture

The most far-reaching implication of FVNST is its mandate for a conceptual re-framing of agricultural vegetation design. It elevates the task from a peripheral, protective activity, i.e., a conservation biology appendix, to farming, to the central architectural core of an agroecosystem's infrastructure. Crop fields are no longer the sole unit of production, but merely one specialized functional module, the "Yield Module", within a larger, engineered functional vegetation network. The performance of the Yield Module is directly dependent on the integrity of its supporting Nourishment, Shelter, and Connectivity Modules. This re-framing aligns agricultural logic with the principles of spatial resilience (Cumming, 2011) and ecosystem service generation, where the service provider is not an organism but the organized network itself (Zhang, 2027b). Crucially, the mathematical structure of FVNST, with its defined thresholds and synergistic functions, provides a computable ecological kernel. This opens the door for the development of AI-assisted decision-support systems that can ingest landscape geodata, optimize the placement of functional modules to maximize the FVNEI, and provide farmers with a precision-designed blueprint for a self-regulating pest control infrastructure.

7 Conclusions and Perspective

This paper has systematically argued that the reliable translation of vegetational diversity into pest suppression is not a problem of adding more diversity, but a problem of engineering a dynamic, threshold-driven ecological network. The Functional Vegetation Network Synergy Theory provides a rigorous, original framework for this transformation. Its core thesis is that sustainable biological control emerges as a system property only when a triadic set of conditions, i.e., irreducible functional module completeness (M), supra-threshold spatial network connectivity (Φ), and an unbroken temporal resource continuum (T), are synergistically co-satisfied. The formal PSP model captures this non-linear emergence, providing a tool for prediction and diagnosis.

The immediate empirical task is to transition FVNST from a logical construct to a fully validated, quantitative theory through a program of long-term field studies. These must be specifically designed to manipulate M , Φ , and T independently and measure the predicted non-linear responses in pest suppression and network robustness. These experiments must be landscape-scale, multi-year works that can capture the slow dynamics of metapopulation establishment and the accumulation of functional network effects. The theory further opens a research frontier at the intersection of network science and ecological field experimentation, pushing for the development of AI-based optimization algorithms that can solve the spatial design problem for FVNEI maximization under socio-economic constraints. The ultimate outlook is the integration of this theoretical kernel into regional and national digital agriculture platforms, and its embedding within agri-environmental policy schemes. Payments for ecosystem services could move from rewarding simple actions, like planting a fixed area of flowers, to rewarding the achievement of a certified functional network integrity standard, verified through a calculated FVNEI. This shift would directly incentivize ecological functionality, not just its superficial appearance, guiding agriculture toward a truly smart, resilient, and self-sustaining future.

References

- Andow DA. 1991. Vegetational diversity and arthropod population response. *Annual Review of Entomology*, 36: 561-586. <https://www.annualreviews.org/content/journals/10.1146/annurev.en.36.010191.003021>
- Altieri MA. 1996. *Agroecology: the Science of Sustainable Agriculture*, 2nd Edition. CRC Press, Boca Raton, USA. <https://www.taylorfrancis.com/books/edit/10.1201/9780429495465/agroecology-miguel-altieri>
- Altieri MA. 1999. The ecological role of biodiversity in agroecosystems. *Agriculture, Ecosystems and Environment*, 74(1-3): 19-31. <https://www.sciencedirect.com/science/article/pii/S0167880999000286>
- Altieri MA, Letourneau DK, Risch SJ. 1984. Vegetation diversity and insect pest outbreaks. *Critical Reviews in Plant Sciences*, 2(2): 131-169. <https://www.tandfonline.com/doi/abs/10.1080/07352688409382193>
- Altieri MA, Nicholls CI. 2004. *Biodiversity and Pest Management in Agroecosystems*. CRC Press, Boca Raton, USA. <https://www.taylorfrancis.com/books/mono/10.1201/9781482277937/biodiversity-pest-management-agroecosystems-miguel-altieri-clara-nicholls>
- Araj SA, Wratten SD, Lister AJ, Buckley HL. 2009. Floral diversity, parasitoids and hyperparasitoids – A laboratory approach. *Basic and Applied Ecology*, 9(5): 588-597. <https://www.sciencedirect.com/science/article/pii/S1439179107000977>
- Balzan MV, Bocci G, Moonen AC. 2016. Utilisation of plant functional diversity in wildflower strips for the delivery of multiple agroecosystem services. A review. *Entomologia Experimentalis et Applicata*, 158: 304-319. <https://onlinelibrary.wiley.com/doi/10.1111/eea.12403>
- Barbosa P, Hines J, Kaplan I, Martinson H, Szczepaniec A, Szendrei Z. 2009. Associational resistance and associational susceptibility: Having right or wrong neighbors. *Annual Review of Ecology, Evolution, and Systematics*, 40: 1-20. <https://www.annualreviews.org/content/journals/10.1146/annurev.ecolsys.110308.120242>
- Bianchi FJJA, Booij CJH, Tscharntke T. 2006. Sustainable pest regulation in agricultural landscapes: A review on landscape composition, biodiversity and natural pest control. *Proceedings of the Royal Society B*, 273(1595): 1715-1727. <https://royalsocietypublishing.org/doi/10.1098/rspb.2006.3530>
- Blitzer EJ, Dormann CF, Holzschuh A, Klein AM, Rand TA, Tscharntke T. 2012. Spillover of functionally important organisms between managed and natural habitats. *Agriculture, Ecosystems and Environment*, 146(1): 34-43. <https://www.sciencedirect.com/science/article/pii/S0167880911003306>
- Cadotte MW, Carscadden K, Mirotchnick N. 2011. Beyond species: Functional diversity and the maintenance

- of ecological processes and services. *Journal of Applied Ecology*, 48(5): 1079-1087. <https://besjournals.onlinelibrary.wiley.com/doi/full/10.1111/j.1365-2664.2011.02048.x>
- Carpenter SR, Walker BH, Anderies JM, Abel N. 2001. From metaphor to measurement: Resilience of what to what? *Ecosystems*, 4(8): 765-781. <https://scispace.com/pdf/from-metaphor-to-measurement-resilience-of-what-to-what-22md8g3yu4.pdf>
- Carson R. 1962. *Silent Spring*. Houghton Mifflin, USA. <https://www.rachelcarson.org/silent-spring>
- Chaplin-Kramer R, O'Rourke ME, Blitzer EJ, Kremen C. 2011. A meta-analysis of crop pest and natural enemy response to landscape complexity. *Ecology Letters*, 14(9): 922-932. <https://onlinelibrary.wiley.com/doi/full/10.1111/j.1461-0248.2011.01642.x>
- Crawley MJ. 1992. *Natural Enemies: The Population Biology of Predators, Parasites and Diseases*. Blackwell Scientific Publications, Oxford, UK. <https://www.wiley.com/en-mx/shop/general-introductory-life-sciences/natural-enemies-the-population-biology-of-predators-parasites-and-diseases-p-9780632026982>
- Cumming GS. 2011. Spatial resilience: Integrating landscape ecology, resilience, and sustainability. *Landscape Ecology*, 26(7): 899-909. <https://link.springer.com/article/10.1007/s10980-011-9623-1>
- Dassou AG, Tixier P. 2016. Response of pest control by generalist predators to local-scale plant diversity: A meta-analysis. *Ecology and Evolution*, 6(4): 1143-1153. <https://onlinelibrary.wiley.com/doi/10.1002/ece3.1917>
- Dib H, Sauphanor B, Capowiez Y. 2016. Effect of management strategies on arthropod communities in the colonies of rosy apple aphid, *Dysaphis plantaginea* Passerini (Hemiptera: Aphididae) in south-eastern France. *Agriculture, Ecosystems and Environment*, 216: 203-206. <https://www.sciencedirect.com/science/article/pii/S0167880915301110>
- Ehler LE. 2006. Integrated pest management (IPM): Definition, historical development and implementation, and the other IPM. *Pest Management Science*, 62(9): 787-789. <https://scijournals.onlinelibrary.wiley.com/doi/10.1002/ps.1247>
- Elhassanein A. 2014. Complex dynamics of a stochastic discrete modified Leslie-Gower predator-prey model with Michaelis-Menten type prey harvesting. *Computational Ecology and Software*, 4(2): 116-128. [http://www.iaees.org/publications/journals/ces/articles/2014-4\(2\)/dynamics-of-a-stochastic-discrete-modified-Leslie-Gower-model.pdf](http://www.iaees.org/publications/journals/ces/articles/2014-4(2)/dynamics-of-a-stochastic-discrete-modified-Leslie-Gower-model.pdf)
- Elsadany AEA. 2012. Dynamical complexities in a discrete-time food chain. *Computational Ecology and Software*, 2(2): 124-139. [http://www.iaees.org/publications/journals/ces/articles/2012-2\(2\)/dynamical-complexities-in-a-discrete-time-food-chain.pdf](http://www.iaees.org/publications/journals/ces/articles/2012-2(2)/dynamical-complexities-in-a-discrete-time-food-chain.pdf)
- Finke DL, Denno RF. 2005. Predator diversity and the functioning of ecosystems: The role of intraguild predation in dampening trophic cascades. *Ecology Letters*, 8(12): 1299-1306. <https://onlinelibrary.wiley.com/doi/abs/10.1111/j.1461-0248.2005.00832.x>
- Gurr GM, Wratten SD, Landis DA, You MS. 2017. Habitat management to suppress pest populations: Progress and prospects. *Annual Review of Entomology*, 62: 91-109. <https://www.annualreviews.org/content/journals/10.1146/annurev-ento-031616-035050>
- Haaland C, Naisbit RE, Bersier LF. 2011. Sown wildflower strips for insect conservation: A review. *Insect Conservation and Diversity*, 4(1): 60-80. <https://resjournals.onlinelibrary.wiley.com/doi/full/10.1111/j.1752-4598.2010.00098.x>
- Hanski I. 1999. *Metapopulation Ecology*. Oxford University Press, Oxford, UK. <https://academic.oup.com/book/53602?login=false>
- Hanski I, Ovaskainen O. 2000. The metapopulation capacity of a fragmented landscape. *Nature*, 404(6779): 755-758. <https://www.nature.com/articles/35008063>

- Heong KL, Aquino GB, Barrion AT. 1991. Arthropod community structures of rice ecosystems in the Philippines. *Bulletin of Entomological Research*, 81(4): 407-416. <https://www.cambridge.org/core/journals/bulletin-of-entomological-research/article/arthropod-community-structures-of-rice-ecosystems-in-the-philippines/60F68DEEAC66EEA7655620C39644E486>
- Iamba K, Teksep C. 2021. Biological role of marigold (*Tagetes erecta* L.) in habitat manipulation and sustenance of natural enemy populations in upland rice. *Arthropods*, 10(3): 66-81. [http://www.iaees.org/publications/journals/arthropods/articles/2021-10\(3\)/1-Iamba-Abstract.asp](http://www.iaees.org/publications/journals/arthropods/articles/2021-10(3)/1-Iamba-Abstract.asp)
- IRRI. 1998. IRRI 1997-1998: Biodiversity Maintaining the Balance. International Rice Research Institute, Los Banos, Philippines. https://books.google.com.sg/books/about/IRRI_1997_1998.html?id=g0jyYu6OZT4C&redir_esc=y
- Jiang LQ, Zhang WJ. 2015. Effects of parasitism on robustness of food webs. *Selforganizology*, 2(2): 21-34. [http://www.iaees.org/publications/journals/selforganizology/articles/2015-2\(2\)/effects-of-parasitism-onrobustness-of-food-webs.pdf](http://www.iaees.org/publications/journals/selforganizology/articles/2015-2(2)/effects-of-parasitism-onrobustness-of-food-webs.pdf)
- Jonsson M, Wratten SD, Landis DA, Gurr GM. 2008. Recent advances in conservation biological control of arthropods by arthropods. *Biological Control*, 45(2): 172-175. <https://www.sciencedirect.com/science/article/pii/S104996440800008X>
- Krause AE, Frank KA, Mason DM, Ulanowicz RE, Taylor WW. 2003. Compartments revealed in food-web structure. *Nature*, 426(6964): 282-285. <https://www.nature.com/articles/nature02115>
- Khan MS, Khan MA, Shabbir MS, et al. 2019. Stability, bifurcation and chaos control in a discrete-time prey-predator model with Holling type-II response. *Network Biology*, 9(3): 58-77. [http://www.iaees.org/publications/journals/nb/articles/2019-9\(3\)/stability-bifurcation-and-chaos-control-in-a-prey-predator-model.pdf](http://www.iaees.org/publications/journals/nb/articles/2019-9(3)/stability-bifurcation-and-chaos-control-in-a-prey-predator-model.pdf)
- Kleiman BM, Koptur S, Jayachandran K. 2021. Weeds enhance pollinator diversity and fruit yield in mango. *Insects*, 12(12): 1114. <https://doi.org/10.3390/insects12121114>
- Kremen C, Miles A. 2012. Ecosystem services in biologically diversified versus conventional farming systems: Benefits, externalities, and trade-offs. *Ecology and Society*, 17(4): 40. <https://www.ecologyandsociety.org/vol17/iss4/art40/>
- Landis DA, Wratten SD, Gurr GM. 2000. Habitat management to conserve natural enemies of arthropod pests in agriculture. *Annual Review of Entomology*, 45: 175-201. <https://www.annualreviews.org/content/journals/10.1146/annurev.ento.45.1.175>
- Letourneau DK. 1987. The enemies hypothesis: Tritrophic interactions and vegetational diversity in tropical agroecosystems. *Ecology*, 68(5): 1616-1622. <https://esajournals.onlinelibrary.wiley.com/doi/abs/10.2307/1939853>
- Letourneau DK, Armbrrecht I, Rivera BS, Lerma JM, Carmona EJ, et al. 2011. Does plant diversity benefit agroecosystems? A synthetic review. *Ecological Applications*, 21(1): 9-21. <https://esajournals.onlinelibrary.wiley.com/doi/full/10.1890/09-2026.1>
- MacArthur RH, Wilson EO. 1967. *The Theory of Island Biogeography*. Princeton University Press, Princeton, USA. <https://press.princeton.edu/books/paperback/9780691088365/the-theory-of-island-biogeography>
- MacLeod A, Wratten SD, Sotherton NW, Thomas MB. 2004. Beetle banks as refuges for beneficial arthropods in farmland: Long-term changes in predator communities and habitat. *Agricultural and Forest Entomology*, 6(2): 147-154. <https://resjournals.onlinelibrary.wiley.com/doi/10.1111/j.1461-9563.2004.00215.x>
- Martin EA, Dainese M, Clough Y, et al. 2019. The interplay of landscape composition and configuration: New pathways to manage functional biodiversity and agroecosystem services across Europe. *Ecology Letters*, 22(7): 1083-1094. <https://onlinelibrary.wiley.com/doi/10.1111/ele.13265>

- May RM. 1974. *Stability and Complexity in Model Ecosystems*. Princeton University Press, Princeton, USA.
<https://www.jstor.org/stable/j.ctvs32rq4>
- McCann KS. 2000. The diversity-stability debate. *Nature*, 405(6783): 228-233.
<https://www.nature.com/articles/35012234>
- Meehan TD, Werling BP, Landis DA, Gratton C. 2011. Agricultural landscape simplification and insecticide use in the Midwestern United States. *Proceedings of the National Academy of Sciences of USA*, 108(28): 11500-11505. <https://www.pnas.org/doi/10.1073/pnas.1100751108>
- Naylor RL, Ehrlich PR. 1997. Natural pest control services and agriculture. In: *Nature's Services: Societal Dependence On Natural Ecosystems* (Daily GC, ed). 151-174, Island Press, Washington DC, USA.
https://www.jstor.org/stable/jj.41003514?turn_away=true
- Neutel AM, Heesterbeek JAP, de Ruiter PC. 2002. Stability in real food webs: Weak links in long loops. *Science*, 296(5570): 1120-1123. <https://www.science.org/doi/10.1126/science.1068326>
- Nicholls CI, Altieri MA. 2013. Plant biodiversity enhances bees and other insect pollinators in agroecosystems. A review. *Agronomy for Sustainable Development*, 33(2): 257-274. https://food.berkeley.edu/wp-content/uploads/2020/07/2012_Nicholls-and-Altieri_Plant-biodiversity-enhances-bees-and-other-insect-pollinators-in-agroecosystems.-A-review.pdf
- Paredes D, Cayuela L, Gurr GM, Campos M. 2015. Is ground cover vegetation an effective biological control enhancement strategy against olive pests? *PLoS ONE*, 10(2): e0117265.
<https://journals.plos.org/plosone/article?id=10.1371/journal.pone.0117265>
- Parolin P, Bresch C, Desneux N, et al. 2012. Secondary plants used in biological control: A review. *International Journal of Pest Management*, 58(2): 91-100.
<https://www.tandfonline.com/doi/full/10.1080/09670874.2012.659229>
- Pascual-Hortal L, Saura S. 2006. Comparison and development of new graph-based landscape connectivity indices: towards the prioritization of habitat patches and corridors for conservation. *Landscape Ecology*, 21(7): 959-967. <https://link.springer.com/article/10.1007/s10980-006-0013-z>
- Peshin R, Bandral RS, Zhang WJ, et al. 2009. Integrated Pest Management: A Global Overview of History, Programs and Adoption. In: *Integrated Pest Management: Innovation-Development Process* (Vol. 1) (Peshin R, Dhawan AK, eds). 1-50, Springer, Netherlands. <https://scispace.com/pdf/integrated-pest-management-innovation-development-process-4ytg8uu3nl.pdf>
- Pimentel D, Stachow U, Takacs DA, et al. 1992. Conserving biological diversity in agricultural/forestry systems. *BioScience*, 42(5): 354-362. <https://discovery.researcher.life/article/conserving-biological-diversity-in-agricultural-forestry-systems/28dbc1178b1d39d19314aa9ed7cd7a11>
- Power AG. 2010. Ecosystem services and agriculture: tradeoffs and synergies. *Philosophical Transactions of the Royal Society B*, 365(1554): 2959-2971.
<https://royalsocietypublishing.org/rstb/article/365/1554/2959/21324/Ecosystem-services-and-agriculture-tradeoffs-and>
- Qi YH. 2003. *Habitat Heterogeneity and Pest Diffusion in Different Habitats: Models and Algorithms*. MSc Thesis. Sun Yat-sen University, China. <https://d.wanfangdata.com.cn/thesis/Y505833>
- Rana SMS. 2015. Chaotic dynamics in a discrete-time predator-prey food chain. *Computational Ecology and Software*, 5(1): 28-47. [http://www.iaees.org/publications/journals/ces/articles/2015-5\(1\)/chaotic-dynamics-in-discrete-time-predator-prey-food-chain.pdf](http://www.iaees.org/publications/journals/ces/articles/2015-5(1)/chaotic-dynamics-in-discrete-time-predator-prey-food-chain.pdf)
- Rand TA, Tylianakis JM, Tschamntke T. 2006. Spillover edge effects: The dispersal of agriculturally subsidized insect natural enemies into adjacent natural habitats. *Ecology Letters*, 9(5): 603-614.
<https://onlinelibrary.wiley.com/doi/epdf/10.1111/j.1461-0248.2006.00911.x>

- Ranjith KG, Kalyan D, Lakshminarayan K, et al. 2019. Crowding effects and depletion mechanisms for population regulation in prey-predator intraspecific competition model. *Computational Ecology and Software*, 9(1):19-36. [http://www.iaees.org/publications/journals/ces/articles/2019-9\(1\)/crowding-effects-and-depletion-mechanisms-of-population-regulation.pdf](http://www.iaees.org/publications/journals/ces/articles/2019-9(1)/crowding-effects-and-depletion-mechanisms-of-population-regulation.pdf)
- Rauwald KS, Ives AR. 2001. Biological control in disturbed agricultural systems and the rapid recovery of parasitoid populations. *Ecological Applications*, 11: 1224-1234. <https://esajournals.onlinelibrary.wiley.com/doi/full/10.1890/1051-0761%282001%29011%5B1224%3ABCIDAS%5D2.0.CO%3B2>
- Risch SJ. 1981. Insect herbivore abundance in tropical monocultures and polycultures: An experimental test of two hypotheses. *Ecology*, 62(5): 1325-1340. <https://esajournals.onlinelibrary.wiley.com/doi/10.2307/1937296>
- Root RB. 1973. Organization of a plant-arthropod association in simple and diverse habitats: The fauna of collards (Brassica oleracea). *Ecological Monographs*, 43(1): 95-124. <https://esajournals.onlinelibrary.wiley.com/doi/abs/10.2307/1942161>
- Rosenheim JA, Kaya HK, Ehler LE, Marois JJ, Jaffee BA. 1995. Intraguild predation among biological-control agents: Theory and evidence. *Biological Control*, 5(3): 303-335. <https://www.sciencedirect.com/science/article/abs/pii/S1049964485710389>
- Rusch A, Chaplin-Kramer R, Gardiner MM, et al. 2016. Agricultural landscape simplification reduces natural pest control: A quantitative synthesis. *Agriculture, Ecosystems and Environment*, 221: 198-204. <https://www.sciencedirect.com/science/article/pii/S0167880916300512>
- Russell EP. 1989. Enemies hypothesis: A review of the effect of vegetational diversity on predatory insects and parasitoids. *Environmental Entomology*, 18(4): 590-599. <https://academic.oup.com/ee/article-abstract/18/4/590/516298?login=false>
- Schellhorn NA, Parry HR, Macfadyen S, Wang Y, Zalucki MP. 2015. Connecting scales: Achieving in-field pest control from areawide and landscape ecology studies. *Insect Science*, 22(1): 35-51. <https://onlinelibrary.wiley.com/doi/full/10.1111/1744-7917.12161>
- Schoenly K, Justo H Jr, Barrion AT, Harris MK, Bottrell DG. 1998. Analysis of invertebrate biodiversity in a Philippine farmer's irrigated rice field. *Environmental Entomology*, 27(5): 1125-1136. <https://academic.oup.com/ee/article-abstract/27/5/1125/2395128?redirectedFrom=fulltext>
- Settle WH, Ariawan H, Astuti ET, Cahyana W, et al. 1996. Managing tropical rice pests through conservation of generalist natural enemies and alternative prey. *Ecology*, 77(7): 1975-1988. <https://repository.ipb.ac.id/handle/123456789/30040>
- Shelton AM, Badenes-Perez FR. 2006. Concepts and applications of trap cropping in pest management. *Annual Review of Entomology*, 51: 285-308. <https://www.annualreviews.org/content/journals/10.1146/annurev.ento.51.110104.150959>
- Shepard BM, Barrion AT, Litsinger JA. 1987. Friends of the Rice Farmer: Helpful Insects, Spiders, and Pathogens. *International Rice Research Institute, Los Banos, Philippines*. https://books.google.com.sg/books/about/Friends_of_the_Rice_Farmer.html?id=m14gAQAAMAAJ&redir_esc=y
- Singh R. 2025. Catalogue of tri-trophic associations of the common hover fly *Ischiodon scutellaris* (Fabricius, 1805) (Syrphidae: Diptera: Insecta) in India. *Arthropods*, 14(4): 225-245. [http://www.iaees.org/publications/journals/arthropods/articles/2025-14\(4\)/6-Singh-Abstract.asp](http://www.iaees.org/publications/journals/arthropods/articles/2025-14(4)/6-Singh-Abstract.asp)
- Stouffer DB, Bascompte J. 2011. Compartmentalization increases food-web persistence. *Proceedings of the National Academy of Sciences*, 108(9): 3648-3652. <https://www.pnas.org/doi/10.1073/pnas.1014353108>

- Straub CS, Finke DL, Snyder WE. 2008. Are the conservation of natural enemy biodiversity and biological control compatible goals? *Biological Control*, 45(2): 225-237. <https://www.sciencedirect.com/science/article/pii/S1049964407001351>
- Taylor PD, Fahrig L, Henein K, Merriam G. 1993. Connectivity is a vital element of landscape structure. *Oikos*, 68(3): 571-573. <https://www.scienceopen.com/document?vid=02fdfed8b6b-4f27-b26a-52fe7310ee24>
- Thebault E, Fontaine C. 2010. Stability of ecological communities and the architecture of mutualistic and trophic networks. *Science*, 329(5993): 853-856. <https://www.science.org/doi/10.1126/science.1188321>
- Tilman D, Cassman KG, Matson PA, Naylor R, Polasky S. 2002. Agricultural sustainability and intensive production practices. *Nature*, 418(6898): 671-677. <https://www.nature.com/articles/nature01014>
- Tiwari KM, Tripathi PN, Singh R. 2024. Biodiversity of aphidophagous predators in Ayodhya district, Uttar Pradesh, India. *Arthropods*, 13(3): 120-125. [http://www.iaees.org/publications/journals/arthropods/articles/2024-13\(3\)/biodiversity-of-aphidophagous-predators.pdf](http://www.iaees.org/publications/journals/arthropods/articles/2024-13(3)/biodiversity-of-aphidophagous-predators.pdf)
- Tscharntke T, Clough Y, Wanger TC, Jackson L, Motzke I, et al. 2012. Global food security, biodiversity conservation and the future of agricultural intensification. *Biological Conservation*, 151(1): 53-59. <https://www.sciencedirect.com/science/article/pii/S0006320712000821>
- Tscharntke T, Klein AM, Kruess A, Steffan-Dewenter I, Thies C. 2005. Landscape perspectives on agricultural intensification and biodiversity – ecosystem service management. *Ecology Letters*, 8(8): 857-874. <https://onlinelibrary.wiley.com/doi/full/10.1111/j.1461-0248.2005.00782.x>
- Tschumi M, Albrecht M, Entling MH, Jacot K. 2015. High effectiveness of tailored flower strips in reducing pests and crop plant damage. *Proceedings of the Royal Society B*, 282(1814): 20151369. <https://royalsocietypublishing.org/rspb/article/282/1814/20151369/84298/High-effectiveness-of-tailored-flower-strips-in>
- Tylianakis JM, Didham RK, Wratten SD. 2004. Improved fitness of aphid parasitoids receiving resource subsidies. *Ecology*, 85(3): 658-666. <https://esajournals.onlinelibrary.wiley.com/doi/full/10.1890/03-0222>
- van Lenteren JC, Bolckmans K, Köhl J, Ravensberg WJ, Urbaneja A. 2018. Biological control using invertebrates and microorganisms: Plenty of new opportunities. *BioControl*, 63(1): 39-59. <https://link.springer.com/article/10.1007/s10526-017-9801-4>
- Wäckers FL. 2004. Assessing the suitability of flowering herbs as parasitoid food sources: Flower attractiveness and nectar accessibility. *Biological Control*, 29(3): 307-314. <https://www.sciencedirect.com/science/article/pii/S1049964403001816>
- Wäckers FL, Romeis J, van Rijn P. 2007. Nectar and pollen feeding by insect herbivores and implications for multitrophic interactions. *Annual Review of Entomology*, 52: 301-323. <https://www.annualreviews.org/content/journals/10.1146/annurev.ento.52.110405.091352>
- Wäckers FL, van Rijn PCJ. 2012. Pick and mix: Selecting flowering plants to meet the requirements of target biological control insects. In: *Biodiversity and Insect Pests: Key Issues for Sustainable Management* (Gurr GM, Wratten SD, Snyder WE, eds). 139-165, Wiley-Blackwell, UK. <https://onlinelibrary.wiley.com/doi/10.1002/9781118231838.ch9>
- Walston L, Hartmann HM, Fox L, et al. 2024. If you build it, will they come? Insect community responses to habitat establishment at solar energy facilities in Minnesota, USA. *Environmental Research Letters*, 19(1): 014053. <https://iopscience.iop.org/article/10.1088/1748-9326/ad0f72>
- Way MJ, Heong KL. 1994. The role of biodiversity in the dynamics and management of insect pests of tropical irrigated rice-a review. *Bulletin of Entomological Research*, 84(4): 567-587.

- <https://www.cambridge.org/core/journals/bulletin-of-entomological-research/article/role-of-biodiversity-in-the-dynamics-and-management-of-insect-pests-of-tropical-irrigated-ricea-review/D604C65C6207DE14990B3B7AD07B5110>
- Zhang WJ. 2007. Computer inference of network of ecological interactions from sampling data. *Environmental Monitoring and Assessment*, 124: 253-261. <https://link.springer.com/article/10.1007/s10661-006-9223-8>
- Zhang WJ. 2011. Constructing ecological interaction networks by correlation analysis: hints from community sampling. *Network Biology*, 1(2): 81-98. [http://www.iaees.org/publications/journals/nb/articles/2011-1\(2\)/Constructing-ecological-interactionnetworks-by-correlation-analysis.pdf](http://www.iaees.org/publications/journals/nb/articles/2011-1(2)/Constructing-ecological-interactionnetworks-by-correlation-analysis.pdf)
- Zhang WJ. 2012a. *Computational Ecology: Graphs, Networks and Agent-based Modeling*. World Scientific, Singapore. <https://www.worldscientific.com/worldscibooks/10.1142/8118#t=aboutBook>
- Zhang WJ. 2012b. How to construct the statistic network? An association network of herbaceous plants constructed from field sampling. *Network Biology*, 2(2): 57-68. [http://www.iaees.org/publications/journals/nb/articles/2012-2\(2\)/how-to-construct-the-statistic-network.pdf](http://www.iaees.org/publications/journals/nb/articles/2012-2(2)/how-to-construct-the-statistic-network.pdf)
- Zhang WJ. 2016a. Network robustness: Implication, formulization and exploitation. *Network Biology*, 6(4): 75-85. [http://www.iaees.org/publications/journals/nb/articles/2016-6\(4\)/network-robustness-implication-formulization-and-exploitation.pdf](http://www.iaees.org/publications/journals/nb/articles/2016-6(4)/network-robustness-implication-formulization-and-exploitation.pdf)
- Zhang WJ. 2016b. *Selforganizology: The Science of Self-Organization*. World Scientific, Singapore. <https://www.worldscientific.com/worldscibooks/10.1142/9685>
- Zhang WJ. 2016c. A method for identifying hierarchical sub-networks / modules and weighting network links based on their similarity in sub-network / module affiliation. *Network Pharmacology*, 1(2): 54-65. [http://www.iaees.org/publications/journals/np/articles/2016-1\(2\)/identifying-hierarchical-sub-networks-modules.pdf](http://www.iaees.org/publications/journals/np/articles/2016-1(2)/identifying-hierarchical-sub-networks-modules.pdf)
- Zhang WJ. 2016d. Detecting connectedness of network: A Matlab program and application in tumor pathways and a phylogenic network. *Selforganizology*, 3(4): 117-120. [http://www.iaees.org/publications/journals/selforganizology/articles/2016-3\(4\)/detecting-connectedness-of-network.pdf](http://www.iaees.org/publications/journals/selforganizology/articles/2016-3(4)/detecting-connectedness-of-network.pdf)
- Zhang WJ. 2017. Finding minimum cost flow in the network: A Matlab program and application. *Selforganizology*, 4(2): 30-34. [http://www.iaees.org/publications/journals/selforganizology/articles/2017-4\(2\)/finding-minimum-cost-flow-in-the-network.pdf](http://www.iaees.org/publications/journals/selforganizology/articles/2017-4(2)/finding-minimum-cost-flow-in-the-network.pdf)
- Zhang WJ. 2018a. *Fundamentals of Network Biology*. World Scientific Europe, London, UK. <https://www.worldscientific.com/worldscibooks/10.1142/q0149>
- Zhang WJ. 2018b. Global pesticide use: Profile, trend, cost / benefit and more. *Proceedings of the International Academy of Ecology and Environmental Sciences*, 8(1): 1-27. [http://www.iaees.org/publications/journals/piaees/articles/2018-8\(1\)/global-pesticide-use-Profile-trendcost-benefit.pdf](http://www.iaees.org/publications/journals/piaees/articles/2018-8(1)/global-pesticide-use-Profile-trendcost-benefit.pdf)
- Zhang WJ. 2024. MetaAnaly: The platform-independent computational tool for meta-analysis in the paradigm of new statistics. *Network Biology*, 14(2): 187-214. [http://www.iaees.org/publications/journals/nb/articles/2024-14\(2\)/MetaAnaly.htm](http://www.iaees.org/publications/journals/nb/articles/2024-14(2)/MetaAnaly.htm)
- Zhang WJ. 2025. Eco-Sustainability assessment of Integrated Pest Management (IPM) techniques: Indicator system and calculator. *Computational Ecology and Software*, 15(3): 99-113. [http://www.iaees.org/publications/journals/ces/articles/2025-15\(3\)/2-Zhang-Abstract.asp](http://www.iaees.org/publications/journals/ces/articles/2025-15(3)/2-Zhang-Abstract.asp)
- Zhang WJ. 2026b. The Niche-Energy-Time triadic synergy hypothesis (NET Hypothesis): A unified explanatory framework for determinants of species diversity. *Computational Ecology and Software*, 16(3):

- 243-272. [http://www.iaees.org/publications/journals/ces/articles/2026-16\(3\)/3-Zhang-Abstract.asp](http://www.iaees.org/publications/journals/ces/articles/2026-16(3)/3-Zhang-Abstract.asp)
- Zhang WJ. 2026c. What triggers explosive radiations and mass extinctions? A unified critical transition framework. *Proceedings of the International Academy of Ecology and Environmental Sciences*, 16(3): 89-122. [http://www.iaees.org/publications/journals/piaees/articles/2026-16\(3\)/2-Zhang-Abstract.asp](http://www.iaees.org/publications/journals/piaees/articles/2026-16(3)/2-Zhang-Abstract.asp)
- Zhang WJ. 2026d. Reflecting Integrated Pest Management and constructing a Dynamic Resilience-based IPM framework. *Arthropods*, 15(4): 183-206. [http://www.iaees.org/publications/journals/arthropods/articles/2026-15\(4\)/2-Zhang-Abstract.asp](http://www.iaees.org/publications/journals/arthropods/articles/2026-15(4)/2-Zhang-Abstract.asp)
- Zhang WJ. 2027a. Decline rate estimations and functional extinction time predictions for major terrestrial insect taxa: An integrated framework of critical network decoupling and multidimensional functional thresholds. *Network Biology*, 17(1): 76-101. [http://www.iaees.org/publications/journals/nb/articles/2027-17\(1\)/4-Zhang-Abstract.asp](http://www.iaees.org/publications/journals/nb/articles/2027-17(1)/4-Zhang-Abstract.asp)
- Zhang WJ. 2027b. Standardized construction of ecosystem service accounting and its embedding into national accounting: From fuzzy valuation to precision measurement. *Computational Ecology and Software*, 17(1): 1-29. [http://www.iaees.org/publications/journals/ces/articles/2027-17\(1\)/1-Zhang-Abstract.asp](http://www.iaees.org/publications/journals/ces/articles/2027-17(1)/1-Zhang-Abstract.asp)
- Zhang WJ, Qi YH, Zhang ZG. 2015. A cellular automaton for population diffusion in the homogeneous rectangular area. *Selforganizology*, 2(1): 13-17. [http://www.iaees.org/publications/journals/selforganizology/articles/2015-2\(1\)/a-cellular-automaton-for-population-diffusion.pdf](http://www.iaees.org/publications/journals/selforganizology/articles/2015-2(1)/a-cellular-automaton-for-population-diffusion.pdf)
- Zhang WJ, Wang R, Zhang DL, Wei W, Chen HD. 2014. Interspecific associations of weed species around rice fields in Pearl River Delta, China: A regional survey. *Selforganizology*, 1(3-4): 143-205. [http://www.iaees.org/publications/journals/selforganizology/articles/2014-1\(3-4\)/interspecific-associationsof-weed-species.pdf](http://www.iaees.org/publications/journals/selforganizology/articles/2014-1(3-4)/interspecific-associationsof-weed-species.pdf)